California State University, Monterey Bay Digital Commons @ CSUMB

AES Theses

Department of Applied Environmental Science

Spring 2020

Surface Conditions Drive Changes In Groundfish Species' Populations Along California Coast

Danielle Marie Fabian California State University, Monterey Bay, dfabian@csumb.edu

Follow this and additional works at: https://digitalcommons.csumb.edu/aes_theses

Recommended Citation

Fabian, Danielle Marie, "Surface Conditions Drive Changes In Groundfish Species' Populations Along California Coast" (2020). *AES Theses*. 1. https://digitalcommons.csumb.edu/aes_theses/1

This Master's Thesis (Open Access) is brought to you for free and open access by the Department of Applied Environmental Science at Digital Commons @ CSUMB. It has been accepted for inclusion in AES Theses by an authorized administrator of Digital Commons @ CSUMB. For more information, please contact digitalcommons@csumb.edu.

SURFACE CONDITIONS DRIVE CHANGES IN GROUNDFISH SPECIES' POPULATIONS ALONG CALIFORNIA COAST

A Thesis

Presented to the

Faculty of the

School of Natural Sciences

California State University Monterey Bay

In Partial Fulfillment

of the Requirements for the Degree

Master of Science

in

Environmental Science

by

Danielle Marie Fabian

Term Completed: Spring 2020

CALIFORNIA STATE UNIVERSITY MONTEREY BAY

The Undersigned Faculty Committee Approves the

Thesis of Danielle Fabian:

SURFACE CONDITIONS DRIVE CHANGES IN GROUNDFISH SPECIES'

POPULATIONS ALONG CALIFORNIA COAST

James Lindholm, Chair California State University Monterey Bay

Alison Haupt California State University Monterey Bay

Ken Smith Monterey Bay Aquarium Research Institute, Moss Landing, California

Christine Huffard Monterey Bay Aquarium Research Institute, Moss Landing, California

Kris Roney, Dean Associate VP for Academic Programs and Dean of Undergraduate and Graduate Studies

11 May 2020

Approval Date

Copyright © 2020

by

Danielle Fabian

All Rights Reserved

DEDICATION

To my family. You guys have been my biggest fans and have supported me tirelessly and selflessly throughout this chapter in my life. I could not have done it without you.

To my friends – you know who you are . Thank you for motivating me to keep going when I was tired through pep talks and words of wisdom, and for lifting my spirits in times of need.

To Jacob Stivers. Thank you for being there for me during the home stretch. You are my rock.

I am forever grateful for all of you.

The fact that a cloud from a minor volcanic eruption in Iceland—a small disturbance in the complex mechanism of life on the Earth—can bring to a standstill the aerial traffic over an entire continent is a reminder of how, with all its power to transform nature, humankind remains just another species on the planet Earth.

Slavoj Žižek

ABSTRACT

Surface Conditions Drive Changes In Groundfish Species' Populations Along California Coast by Danielle Marie Fabian Master of Science in Environmental Science California State University Monterey Bay, 2020

Oceans display physical variability over a range of temporal and spatial scales, influencing factors such as larval dispersal, nutrient availability, species migration, and biodiversity. Such variability is vulnerable to the effects of climate change. Larvae and juveniles are particularly susceptible to changes in ocean variability, and changes in the early life stages of species populations ultimately impacts the adult stages. Physical consequences stemming from climate driven variability, including a loss of key prey species and changes in depth distributions of vulnerable fish species, could ultimately impact ecosystem services and threaten human food security and fisheries. The deep sea benthopelagic groundfish play significant roles in biogeochemical and ecological processes on a global scale. Macrourids, the most common benthopelagic fish in the deep sea, are important benthic and midwater predators; making it critical to understand changes in their population structures over time. Macrourid larvae reside in surface waters, where they are more vulnerable to climate driven variability. Merluccius productus (M. productus) is a commercially important benthopelagic groundfish species with a life history similar to Macrourids. The combination of surface waters exhibiting more changes in physical variability compared to the deep sea and the vulnerability of larvae to oceanographic variability makes observing these benthopelagic species particularly important for understanding the full-depth interactions and climaterelated changes. This project focuses on studying correlations between changes in the life stages of Macrourid species' and M. productus' populations over time in relation to changes in climate and surface ocean conditions. Four research questions are addressed; Q1 & Q2) Has groundfish (specifically Macrourid spp. (Q1) and M. productus (Q2)) population structure changed over time in relation to changing climate and surface-ocean conditions?; and Q3 & Q4) Has the depth range of groundfish species (specifically Macrourid spp. (Q3) and M. productus (Q4)) changed over time with changing climate and surface-ocean conditions? With increasing changes in climate and surface-ocean conditions over time, there is a significant change in both Macrourid species' population structure and M. productus population structure over time (Q1 and Q2). With increasing changes in climate and surfaceocean conditions over time, there is no significant change in M. productus' depth distribution over time (Q4). Regarding potential changes in Macrourid species' depth distributions, Q3 couldn't be answered due to insufficient data. This study increases our understanding of how potential impacts of changing climate and surface-ocean conditions on M. productus' population structure and depth distribution could affect Macrourid populations. Findings also provide valuable insight for: 1) predicting and managing consequences of climate change and 2) management of coastal and marine resources and fisheries in the future.

TABLE OF CONTENTS

vii

vi
ix
x
xi
1
1
6
9
9
11
11
11
12
13
14
14
14
16
17
18
19
20
21

4 DISCUSSION & CONCLUSIONS	
Macrourid & Merluccius productus Observations (2011-2017)	27
Spatial Scale & Location Significance	
Potential Impacts of Climate and Fishing Pressure	
REFERENCES	
A LIST OF HYPOTHESIS	51
B RESULTS TABLES	56
C RESULTS PLOTS	59

LIST OF TABLES

PAGE

Table 1. Table of statistical analyses results for relationships between Macrourid spp.'	
variables and environmental variables.	57
Table 2. Table of statistical analyses results for relationships between <i>M. productus</i> '	
variables and environmental variables.	58

LIST OF FIGURES

PAGE

Figure 1. Focal species	8
Figure 2. Station M. (Image credit: MBARI)	11
Figure 3. Instruments currently deployed at Sta. M (Image credit; MBARI)	11
Figure 4. Time-lapse camera tripod image	13

ACKNOWLEDGEMENTS

This work was supported through a collaboration with the Monterey Bay Aquarium Research Institute, which is funded by the David and Lucile Packard Foundation. I'd like to give special thanks to the members on my thesis committee, including my advisor, James Lindholm, and Ken Smith, my mentor at MBARI; as well as Alison Haupt and Christine Huffard. I would also like to thank Linda Kuhnz, Henry Ruhl, Larissa Lemon, Paul McGill, Waldo Wakefield, Ian Taylor, Ian Stewart, George Matsumoto, the Smith lab, and MBARI staff for their help with sample collection and analysis. Finally, I would like to thank Susan Alexander, Alyssa McNally, my professors, and other CSUMB staff for their guidance along the way of completing my thesis work.

CHAPTER 1

INTRODUCTION

CLIMATE CHANGE EFFECTS ON ENVIRONMENTAL VARIABILITY & FISHERIES

Deep sea groundfish along continental margins play vital roles within their ecosystems (Hamilton et al. 2004), including major biogeochemical cycles (Ragueneau et al. 2000; Street-Perrott et al. 2008), acting as an important process contributing to marine primary production and carbon export to the deep sea (Tréguer and De La Rocha 2013; Schneider et al. 2008). The eastern division of the North Pacific is highly biodiverse, containing a combination of subarctic, temperate, and subtropical groundfish species (Therriault et al., 2016). An east to west diversity gradient is present there, with higher diversity and the majority of fish biomass occurring within coastal nearshore areas and the continental shelf, compared to the less diverse oceanic realm (Mueter and Norcross, 2002). Due to the great depths that deep sea groundfish reside in, the data needed to understand and monitor these species is lacking (Therriault et al., 2016). Grenadiers, of the family Macrouridae (Neat and Burns 2010), are the most common benthopelagic groundfish in the deep sea (Jeffreys et al. 2010; Matsui et al. 1991). For the sake of consistency, "grenadiers" will be referred to as Macrourids for the entirety of this paper. Macrourids are important benthic and midwater predators (Morley & Belchier, 2002) and some species are targeted by fisheries. Many Macrourids, including four of the five focal species of this research project, are benthopelagic (Hiroaki et al. 2002). The adult stages reside in the deep sea, inhabiting the water just above the seafloor, while the larvae are planktonic and reside in the upper water column (Allen et al. 2006). The pelagic larvae is ultimately the link between the surface water and their adult habitat – the seafloor (e.g. Stein & Pearcy, 1982). The extent to which Macrourid, and other deep sea benthopelagic groundfish, populations are coupled to surface conditions, and the role that different marine species interactions play within this coupling are examples of important unresolved questions in the field of marine ecology that need to be addressed regarding these species.

Oceans display physical variability over a range of temporal and spatial scales (Brierley and Kingsford 2009; Fraschetti et al. 2005; Powell et al. 2006), influencing factors such as larval dispersal (Botsford et al. 1994), physiology (Hinrichsen 2009), nutrient availability (Kamykowski et al. 2002), species migration (Wegner et al. 2003), production, and biodiversity (Lindegren et al. 2016). In California waters, groundfish species' populations experience different degrees of physical variability interannually, depending on the depth range and relative spatial location of groundfish species' populations residing in the California Current Ecosystem (CCE) (Bograd et al., 2019). The combination of three different source waters and various currents within the California Current Ecosystem allows for great spatial diversity in physical and biological processes (Strub et al., 1990; Mendelssohn et al., 2003). Species residing in regions further north and closer to the coast experience more extreme temperature variation due to the cooling of waters, both at the surface and at depth, from the intrusion of Subarctic water and upwelling in the spring and summer (Mendelssohn et al., 2003; King et al., 2011). By contrast, annual thermocline variability in the southern and offshore regions is very low, partially due to subtropical water dominating these regions and stratification reduction in the seasonal mixed layer during the summer and autumn months (King et al., 2011).

Similarly, surface waters exhibit more changes in physical variability compared to the deep sea (McLeod et al. 2009; Davies et al. 2007). Shallower living groundfish species, such as *Merluccius productus (M. productus)* may experience more extreme physical variability compared to groundfish living deeper in the water column, such as *Coryphaenoides armatus* (*C. armatus*); and therefore, may are more directly affected by such variability (Peterson et al., 2006; Bailey et al., 1982). However, changes in the surface waters ultimately affect deep sea ecosystems, impacting deep sea species either directly or indirectly (Kuhnz et al., 2014; Moore et al., 2013). For instance, extended periods of high-nutrient conditions caused by upwelling result in high productivity in surface waters (Pickett & Paduan, 2003; Bograd et al., 2019), which eventually leads to increased food supply reaching the deep sea in the form of particulate organic carbon (Buesseler et al., 2008; Steinberg et al., 2012). Additionally, *M. productus* carrion has been found to be the largest food source for various groundfish species, including Macrourids (Drazen et al., 2009; Dufault et al., 2009), making up 61% of

pelagic biomass in the California Current ecosystem (CCE) (Agostini et al., 2006; Drazen et al., 2012). Although physical variability in the surface waters may not directly affect deep sea Macrourids, these species are dependent on the carrion of *M. productus* ' populations, which are more sensitive to such variability (Agostini, 2005). For instance, the location of spawning can shift slightly with temperature (Agostini, 2005). Direct physical variability effects on *M. productus* populations could in turn indirectly affect Macrourids and other deep sea groundfish species . The extent to which physical variability affects deep sea benthopelagic species, both directly and indirectly, are two important and unresolved questions that need to be addressed in the field of marine ecology.

Physical ocean variability among and within all marine ecosystems is vulnerable to the effects of climate change (Brierley and Kingsford 2009; Doney et al. 2012), primarily due to temperature increases (Rosenzweig et al., 2008). Since the mid-1800s, there has been a tremendous increase in carbon dioxide (CO_2) in the atmosphere due to the combustion of fossil fuels on a global scale (Karnosky 2003; Patz et al. 2014). This rise in C02 aggravates the greenhouse gas effect (Peters 1985; Recent Global Warming 2002), making the planet warmer than it otherwise would be (Recent Global Warming 2002; Mitchell 1989; Schneider 1989). Additional effects of climate change include: rising sea levels, rising global mean seasurface temperatures (SST), increased ocean acidity, perturbed regional weather patterns, altered ocean circulation, changes in the extent of oxygen-deficient dead-zones, and changed nutrient loads (Brierley and Kingsford 2009; Doney et al. 2011). Marine biological processes are affected by such physical consequences, which could ultimately impact ecosystem services and threaten human food security (Brander 2007; Pecl et al. 2017). For example, significant changes in physical ocean variability caused by ocean warming (Rosenzweig et al., 2008) can affect the conditions of larvae and juveniles (Pansch et al. 2012), as well as the timing of reproduction and reproductive output (Velaquez 2003; Pankhurst and Munday 2011; Petrik et al. 2015).

Larvae and juveniles are particularly susceptible to changes in ocean variability (Pansch et al. 2012; Heath 1992), and changes in the early life stages of species populations ultimately impacts the adult stages (Pankhurst and Munday 2011; Heath 1992). Juveniles and

larvae may not be able to survive the changes in salinity, temperature, and pH that are tolerated by adult stages (Rijnsdorp et al. 2009; Byrne and Przeslawski 2013). Similarly, if the hatching time of larval eggs does not coincide with food availability, survival rates may be affected (Brierley and Kingsford 2009; Fiksen and Jorgensen 2011). Changes in food availability could stem from temperature-driven phenological changes that could alter the timing of plankton blooms (Berner et al. 2018); ultimately leading to a loss of food supply for prey species, and potentially causing shifts in prey species distributions (Brierley and Kingsford 2009). The loss of key prey species could negatively impact remaining predators that not only play an important role within their ecosystems, but that are also deemed important in commercial fisheries (Young et al. 2015; Ripple et al. 2014). The combination of surface waters exhibiting more changes in physical variability compared to the deep sea (McLeod et al. 2009; Davies et al. 2007) and the vulnerability of larvae to oceanographic variability change (biological and physical) (Pansch et al. 2012; Heath 1992) makes observing deep sea benthopelagic species particularly important for understanding the interactions and feedbacks between marine systems and climate-related changes. Changes in surface water conditions probably affect larval stages of these species. Although adult fishes living in the deep sea may experience fewer negative effects caused by changing climate and surface conditions than their pelagic larvae and juveniles, effects on the early life stages will ultimately affect the adult populations and the fisheries dependent upon them (Johnson et al. 2011). In addition, direct impacts of increasing changes in climate and surface-ocean conditions on the more sensitive shallower species may potentially affect deep sea benthopelagic species indirectly.

Another concern is that extremely vulnerable fish species have exhibited changes in depth distribution relating to changing water conditions (Nye et al. 2009; Cheung et al. 2010). Fishing effort is venturing to deeper depths with increasing depletion of economically important species in shallower water (Norse et al. 2012; Yoklavich et al. 2000), and changes in depth distributions can potentially impact the risk of individuals being caught as bycatch (Gilly et al. 2013; Roberts 2002). In addition, Macrourids generally have slow recovery rates from fishing mortality (Norse et al., 2012), due to high longevity and slow growth rates (Clark, 2001), and Macrourid species that are commercially exploited may be overfished

(Shotton 2005). The combination of slow recovery rates and overfishing could potentially result in a fishery collapse within a short time period (Large et al. 2003; Shotton 2005). A good example for showing the importance of monitoring Macrourid species' populations can be seen in reference to the commercially targeted species, Coryphaenoides acrolepis (C. acrolepis), and the non-target species, Coryphaenoides armatus (C. armatus). These two species overlap on the slope. *Coryphaenoides armatus*, although of little interest to fisheries, is often caught as bycatch and is difficult to differentiate from C. acrolepis (Iwamoto et al. 2015). Both C. acrolepis and C. armatus are known to have pelagic larvae (e.g. Stein & Pearcy, 1982). It is important to assess if there is any change in C. acrolepis and C. armatus pelagic larval density and benthopelagic adult fishes with the change in climate and surface conditions in order to provide insight necessary to manage fisheries in the future. Accidental misclassification could result in fishermen wrongly assuming that the target species', C. acrolepis, population is thriving. In reality, it might be declining at a rapid rate. This could result in overexploitation and collapse of fisheries, which in turn, could negatively affect local communities as a whole. Misidentification could also lead to overexploitation of the non-target species, potentially disrupting deep-sea ecosystems (Gjerde 2006). For instance, declining groundfish populations, resulting from anthropogenic climate change and overfishing, could potentially create shifts in marine phytoplankton communities (Pershing et al. 2015; Katz et al. 2009).

Given the extent to which deep sea benthopelagic groundfish are coupled to surface conditions and the vulnerability of their early life stages to changes in physical variability, the goal of this thesis is to better understand, how increasing changes in climate and surfaceocean conditions over time are affecting deep sea benthopelagic groundfish species off the coast of California. To achieve this, this study monitors changes occurring in larval and adult life stages of environmentally and economically important benthopelagic fish species with changes in climate and surface-ocean conditions. Findings from this study have the potential to provide valuable insight for the global climate change discussion in that it sheds light on both the direct effects and potential indirect effects of changing climate and surface conditions on economically and ecologically important benthopelagic groundfish species. In turn, such insight can be incorporated into the management of coastal and marine resources to aid in the protection of both fisheries and the marine ecosystems they depend on.

FOCAL SPECIES

The five focal species in this study are: *C. armatus, Coryphaenoides leptolepis* (*C. leptolepis*), *Coryphaenoides yaquinae* (*C. yaquinae*), *C. acrolepis*, and *M. productus*, commonly known as North Pacific hake. For the sake of consistency, Pacific hake will be referred to as *M. productus* for the entirety of this thesis. All of the *Coryphaenoides* species are included in the family Macrouridae (Cohen et al. 1990). Although not a Macrourid, *M. productus* is a benthopelagic groundfish species and is included in the study due to its similar life history (Melnychuk et al. 2013; Cohen et al. 1990). *Merluccius productus* and Macrourids are all included in the order Gadiformes (Iwamoto et al. 2015). For this study, Macrourids were analyzed as a group and *M. productus* was analyzed independently as a single species. Due to difficulty distinguishing *C. armatus*, *C. leptolepis*, and *C. yaquinae* from one another at the long-term time series study site, all three species were grouped together as one complex (*Coryphaenoides armatus-leptolepis-yaquinae* complex or Macrourid species complex) for analyses at this site. The following are brief descriptions of the Macrourid focal species in this project:

Coryphaenoides acrolepis:

Coryphaenoides acrolepis is a benthopelagic mid-slope Macrourid species that is significantly more abundant than any other large fish species living on the continental slope (Cohen et al. 1990). Commercially, this species is the North Pacific's most important Macrourid species (Koslow 1996; Laptikhovsky et al., 2013). It is a non-migratory species that resides in the North Pacific at a depth range of 400-1800 meters (Matsui et al. 1991; Cohen et al. 1990). This species has pelagic larvae (Iwamoto T. 1999). Due to its economic importance, it is necessary to monitor potential changes in abundance of this species over time to avoid overfishing (as recommended for groupers: Mangubhai et al. 2011).

Coryphaenoides armatus:

Coryphaenoides armatus is a deep-slope upper continental rise Macrourid species found in all the world's oceans, at depths between 800 and 4,000 meters (Merrett and Haedrich 1997; Cohen et al. 1990). This species is known to grow slowly, potentially living up to 75 years of age (Cohen et al. 1990). Distribution of *C. armatus* is likely driven by migration in response to variation in food availability (Drazen et al. 2012; Bailey et al. 2006). *Coryphaenoides armatus*, although of no interest to fisheries, is often caught as by-catch and is difficult to differentiate from *C. acrolepis* (Iwamoto et al. 2015). This species has pelagic larvae (Iwamoto T. 1999).

Coryphaenoides yaquinae:

Coryphaenoides yaquinae is an abyssal Macrourid species that is confined to abyssal depths encompassed within the Pacific Ocean (Jamieson et al. 2012). Although preferring deeper depths, the species is known to inhabit a depth range of 3400 to 5800 meters (Wilson and Waples 1983). *Coryphaenoides yaquinae* is closely related to *C. armatus*, and the two species are almost indistinguishable from one another based on physical characteristics alone (Smith et al. 1979; Jamieson et al. 2012). While generally segregated by depth, the two species co-exist on the Pacific continental margin between the 3400-4300 meter depth ranges (Jamieson et al. 2012; Endo and Okamura 1992). This species has pelagic larvae (Iwamoto T. 1999).

Coryphaenoides leptolepis:

Coryphaenoides leptolepis is commonly referred to as the "Ghostly Grenadier" and is listed as Least Concern because the species has no known predators and resides outside the range of fisheries (Iwamoto 2015). It is a bathydemersal species, inhabiting a depth range of 610-4000 meters, and can grow to 62 centimeters in length (Iwamoto 2015). *Coryphaenoides leptolepis* is native to the northeast and eastern central Pacific, as well as the western central, eastern central, northwest, and northeast Atlantic (Wilson and Attia 2003; Iwamoto 2015). Due to a lack of commercial importance and depth range, very little is known about *C. leptolepis* (Iwamoto 2015). This species has pelagic larvae (Iwamoto T. 1999).

Merluccius productus:

Merluccius productus, of the family Merlucciidae, is commonly referred to as the Pacific hake (Carrera et al. 2006). It is a benthopelagic species (Essington et al. 2011) with pelagic larvae (Leaman and Beamish 1984) and is important prey for many larger fishes and marine mammals (Hoff et al. 2015). Within U.S. waters, *M. productus* is considered one of the most important commercial fishery species (NOAA, Pacific whiting (hake)). *Merluccius productus* resides in surface waters and depths up to 1,000 meters and is native to the eastern Pacific (Cohen et al. 1990). It inhabits coastal and oceanic areas but resides mainly on the continental shelf (Alverson and Larkins 1969; Cohen et al. 1990). *Merluccius productus* is a highly migratory species and can grow to 91 centimeters in length (Beamish et al. 2005; Hoff et al. 2015). Currently, this species is the most abundant groundfish population in the California Current system, with more *M. productus* caught than all other groundfish species combined (Fisheries and Oceans Canada 2018).



Figure 1. Panels A-E showing separate images of the benthopelagic groundfish focal species chosen for this study.

ENVIRONMENTAL VARIABLES

In order to monitor potential changes occurring in larval and adult life stages of Macrourids and *M. productus* in relation to changes in climate and surface-ocean conditions, five environmental variables were chosen and compared to Macrourid and M. productus data. Certain variables, despite not being classified as "climate" or "surface-ocean condition" variables, were chosen due to their strong relationships with climate and surface-ocean conditions. The five environmental variables used in the study are the following: sea surface temperature (SST), particulate organic carbon (POC) flux reaching the deep sea, carbon to nitrogen ratio (C:N) of sinking particulate matter, net primary production (NPP), and the Multivariate ENSO index (MEI). NPP, POC and C:N are indicators of food supply (Buesseler et al., 2008; Steinberg et al., 2012), however, they are useful for monitoring climate-related changes because such changes affect NPP, POC, and C:N (Moore et al., 2013), ultimately affecting factors such as export efficiency and the community structures of abyssal and deep-sea ecosystems (Kuhnz et al., 2014; Moore et al., 2013). Climate change is predicted to increase certain phytoplankton C:N ratios (Burkhardt et al., 1999). Increased C:N ratios could result in decreased export efficiency (Burkhardt et al., 1999), likely resulting in less POC reaching the seafloor in deeper waters (Moore et al., 2013)). By contrast, export efficiency is expected to increase due to climate-related declines in primary production, ultimately increasing export efficiency of POC (Arteaga et al., 2018; Chavez et al., 2011). This predicted increase in export efficiency will likely outweigh the declines caused by increased C:N ratios of phytoplankton (Chavez et al., 2011). Regardless, C:N, NPP, and POC are useful in assessing associations between fish populations and climate/surface-ocean conditions (Moore et al., 2013). Throughout the remainder of this paper, the variables relating to "climate and surface-ocean conditions" mentioned above, including indirect indicators of food supply, will be referred to as "environmental variables" for the sake of simplicity.

APPROACH

This project focuses on studying correlations between changes in deep-sea benthopelagic groundfish populations over time in relation to changes in climate and surface ocean conditions, as well as changes in indirect indicators of food supply. Four research questions are addressed in this project; Q1) Has Macrourid population structure (abundance, body length, biomass) changed over time in relation to changing climate and surface-ocean conditions? Q2) Has *Merluccius productus* ' population structure (abundance, body length, biomass) changed over time in relation to changing climate and surface-ocean conditions? Q3) Has the depth range of Macrourid species changed over time with changing climate and surface-ocean conditions? Q4) Has the depth range of *Merluccius productus* changed over time with changing climate and surface-ocean conditions? Q4) Has the depth range of *Merluccius productus* changed over time with changing climate and surface-ocean conditions? Question 1 addresses Macrourids using two separate means: 1) analyzing a time-series of benthic imagery annotations, and 2) analyzing Macrourid data collected from external sources. Questions 2-4 were addressed strictly through mined external data. Questions 2 and 4 address potential changes relating to *M. productus*. Due to separate data collection methods for Q1, there are separate component hypotheses for each method, as well as for each of the various fish variables obtained within the mined data. Further details regarding each question, their component hypotheses, and the population structure variables used are given below (See Appendix A).

This thesis involves a two-part study to address the four research questions. Part 1 of the study addresses Q1. Part 2 of the study addresses all four questions. Part 1 of the study involved conducting an hourly benthic imagery annotation to determine changes in adult Macrourid spp. abundance during a 6-year period at a station in the northeast Pacific (Sta. M). The annotation focused on data collected from 2011 to 2017. For Part 2 of the study, two existing NOAA databases (ERDDAP & FRAM; see methods section) were examined to determine potential changes in *M. productus* and Macrourid population structures/depth-distributions over time (2011-2017). The databases contained data from a continuous 64-year time-series, however, only data from the study period (2011-2017) relating to life stages of *M. productus* and Macrourid species residing along the California coast were used in analyses. Population structure and depth distribution data collected at Sta. M and obtained from NOAA databases were then compared to environmental variables relating to climate and surface-ocean conditions to determine if similar temporal trends occurred along the California coast.

CHAPTER 2

METHODS

STUDY AREA

Station M (Sta. M) is an abyssal station in the northeast Pacific, approximately 220 km west of Point Conception along the central California coast (34° 50'N, 123° 00'W). The station is approximately 4,100 m deep. This time series study is maintained by the Smith lab at the Monterey Bay Aquarium Research Institute (MBARI) and has been monitored consistently since 1989 (Smith et al., 2020).

INSTRUMENTATION

A number of autonomous longterm instruments have been deployed on the seafloor at Sta. M to collect various types of data. The instruments used in part 1 of this study are currently deployed at Sta. M and include 2 sediment traps moored 600 and 50 meters above the seafloor (mab) (Smith et al., 2013), as well as



Figure 2. Station M. (Image Credit: MBARI)





2 benthic time-lapse cameras (Smith et al., 2017). The sediment traps collect particulate

matter that falls through the water column, eventually reaching the seafloor (Smith et al., 2013). The traps collect sinking material at 10-day intervals with a rotating carousel of 21 bottles (Smith et al., 1994). Once recovered, these bottles are taken back to the lab where the contents are analyzed for total and organic carbon, allowing for estimates of POC flux (Smith et al., 2013). The time-lapse cameras took a still image of the seafloor every hour for the duration of this study period (Smith et al., 1993), with gaps for periods when the camera systems were being serviced or did not work (Smith et al., 1994). Continuous deployment periods range from 3 to 12 month time spans. The cameras are mounted on a titanium frame at angles of 1) 35° in the vertical plane and 31° in the horizontal plane (Smith 1993), and the camera lens is situated approximately 2 m above the seafloor (Smith et al., 1994). The fields of view extend to 6.5 m from the tripod frame base and begins at 1.8 m in front of the cameras. An area of the seafloor, approximately 20 m², is illuminated by the two strobes on either side of the camera housings (Lemon, 2018). High-resolution digital imagery is used for deployments since 2007 (Sherman & Smith, 2009. Time-lapse imagery allows for observational estimates of various factors, including species' abundances and spatial patterns of individual organisms (Lemon, 2018). Satellite color imagery is also used to obtain SST and regional surface water chlorophyll, providing the data necessary for estimating NPP in surface waters above Sta. M (Smith et al., 2017). Around Sta. M, a radius of 100 km was used for the daily satellite sensing (Smith et al., 2017). Refer to Smith et al., 2017 for a description of full methods.

IMAGE ANALYSIS (PART 1)

Each hourly image from the time-lapse camera, taken from 2007 to 2017, was reviewed for presence of identifiable Macrourids using MBARI's open-source Video Annotation and Reference System (VARS) software (Schlining & Stout, 2006). VARS allows for annotation at any level of detail, including measurements of individual animals on the sea floor and comments on behavior and color. However, individuals in this study were not measured because they were typically in the water column. The VARS software stores all of the data and annotations within the system (Schlining & Stout, 2006). The lower 75% of each 20 m² image was used as an effective field of view due to the uneven lighting and angle of the tripod images, increasing confidence in the identification of Macrourid species.



Figure 4. Time-lapse camera tripod image. The number of individual Macrourids present is counted for each hourly image (indiv./hr) and averaged weekly.

QUESTION 1(A):

Has Macrourid species complex' population structure changed over time in relation to changing climate and surface-ocean conditions?

Null Hypothesis: With changes in climate and surface-ocean conditions over time, there is no significant change (increase or decrease) in Macrourid species complex' abundance.

Alternative Hypothesis: With changes in climate and surface-ocean conditions over time, there is a significant change (increase or decrease) in Macrourid species complex' abundance.

Counts of all benthopelagic Macrourid species present at Sta. M from 2011-2017 were analyzed to determine changes in Macrourid abundance. The three dominant species present at Sta. M were *C. armatus*, *C. leptolepis*, and *C. yaquinae*. As a proxy for local abundance, counts of individuals per frame were determined. Counts were determined by visually counting the number of individuals present in each image's effective field of

view every hour. Due to difficulty distinguishing the three species from one another in the images, all three species were grouped together as one complex (*Coryphaenoides armatus-leptolepis-yaquinae complex*) when counting the number of individuals present. The number of fish present per hour (i.e. per image) was averaged over each week. Weekly averaged Macrourid count data were compared to the environmental variables to determine potential correlations between changes in fish populations at Sta. M. and climate. As previously stated, the variables used in the study included the following: sea surface temperature (SST), particulate organic carbon flux (POC flux), net primary production (NPP), carbon to nitrogen ratio (C:N) in the sinking particulate matter, and the Multivariate ENSO Index (MEI). Once these variables used in the data analyses were extracted for the entire time-series, they were averaged to obtain weekly averages to compare to weekly averaged fish count data using statistical analyses.

MINING DATA FROM OUTSIDE SOURCES (PART 2)

Existing NOAA databases containing data from a continuous 64-year time-series were examined to obtain Macrourid spp.' population structure and depth-distribution data for all life stages of the species present at Sta. M when data were available. Additionally, similar data for the shallower living *M. productus*, *C. acrolepis*, and all deep-sea Macrourid species residing along the California coast were obtained and included in the analyses when data were available. All extracted data falls within the time period of 2011 to 2017. The two NOAA databases include the data server ERDDAP and the data warehouse FRAM (refer to the Literature Cited section for links to both sources).

QUESTIONS 1(B)-4:

Q1(B) & Q2) Has groundfish (specifically Macrourid spp. (Q1) and M. productus (Q2)) population structure changed over time in relation to changing climate and surface-ocean conditions?

Q3 & Q4) Has the depth range of groundfish species (specifically Macrourid spp. (Q3) and M. productus (Q4)) changed over time with changing climate and surface-ocean conditions?

The collected NOAA groundfish data were also averaged (weekly) and compared to the same physical variables as the fish density data from Sta. M via statistical analyses to address the separate research questions. Comparing early life stage data to adult abundance data provides valuable insight into possible larval or juvenile density decline due to climate and surface conditions. A decline in adult fish populations will reflect this early life-stage density decline and can be predicted in the future. Early life stage data, adult abundance data, and other data relating to population structure obtained from NOAA databases were then compared to the adult fish density data collected at Sta. M for the study period (2011-2017) to determine if there are similar temporal trends in population structure occurring at Sta. M and along the California coast. Values derived from statistical analyses on Sta. M and NOAA mined data were assessed to determine possible and/or similar trends. The mined fish variables used in population structure analyses included the following: 1) regional (CA) Macrourid spp. biomass, 2) regional Macrourid spp. catch per unit effort (CPUE), 3) regional and local (relating to Sta. M) M. productus biomass, 4) regional and local M. productus CPUE, 5) regional Macrourid body length, 6) regional *M. productus* body length, 7) regional Macrourid spp.' larval abundance, and 8) regional *M. productus* ' larval abundance.

Depth-distribution data were also extracted along with the NOAA time-series quantitative data to assess any changes in species' depth distributions over time. Statistical analyses were used to determine any potential relationships between changes in species' depth distributions over time and changes in climate and surface conditions. Depth-distribution data were compared to the same environmental variables as the fish density data from Sta. M in statistical analyses. Findings from statistical analyses were compared between Macrourid spp. and *M. productus* to determine similar trends among the different species regarding depth-distributions. The mined fish variables included in depth-distribution analyses were the following: 1) regional Macrourid spp. depth, and 2) regional and local *M. productus* depth.

STATISTICAL ANALYSES:

All data derived from Sta. M annotations and mined NOAA sources were averaged to obtain weekly averages to compare averaged fish population structure variables using three methods of statistical analyses, including; 1) Man-Kendall Trend Test, 2) Cross correlation (CC) analyses, and 3) Spearman's rho. Weekly averages for all previously stated fish variables were obtained and ranked using Excel. In order to detect any statistically significant trends in the time-series (2011-2017) data, the Mann-Kendall method was applied. Simply put, the Mann-Kendall method was used to establish whether Macrourid and *M. productus* variables were increasing or decreasing consistently over time. This is the first step in answering the study questions because we must first establish possible trends in fish variables before we can determine how climate and surface-ocean conditions are affecting or causing such trends. CC analyses were used to determine the separate relationships and the degree of correlation between changes in fish populations, at Sta. M and along the California coast, and changes in climate and surface ocean conditions over time. CC analyses were run in the statistical program R, and accounted for time lags between variables, as well as measured similarity between the variable series. To evaluate the linear relationship between variables, scatter plots were created to incorporate the shifted weekly variables and fish variables. To evaluate the monotonic relationship between variables (i.e., the strength of association between two variables), as well as for comparison purposes, Spearman's rho was used alongside CC analyses. Lag periods, correlation strengths (+, -), and Spearman's rho values, derived from CC analyses on Sta. M and mined data, were assessed to determine possible trends over time. In other words, CC analyses and Spearman's rho were used to answer the remaining portions of this study questions, by determining relationships (and relationship strengths) between Macrourid and *M. productus* population structure variables and variables relating to climate and surface-ocean conditions. CC analyses also allowed us to determine the lag times of the relationships, i.e., the amount of time it takes for fish variables to change in relation to changes in the associated environmental variables.

CHAPTER 3

RESULTS

Findings from this study provided answers to the four research questions addressed in this thesis. These questions are relisted below. Page numbers for the complete list of research questions, their associated component hypotheses, and statistical analyses results tables can be found in the Table of Contents. Three of the fourteen fish variables used in the Mann-Kendall analyses showed significant trends over the timeseries (2011-2017). Macrourid length and abundance increased significantly over the 7-year period (Macr. spp. comp.' abundance: p < 0.0001; Macr. length: p = 0.0023). Similarly, local *Merluccius productus* biomass significantly increased over time (*M. productus* biomass: p = 0.0170).

Population Structure:

Question 1:

HAS MACROURID SPECIES' POPULATION STRUCTURE CHANGED OVER TIME IN RELATION TO CHANGING CLIMATE AND SURFACE-OCEAN CONDITIONS?

QUESTION 2:

HAS MERLUCCIUS PRODUCTUS' POPULATION STRUCTURE CHANGED OVER TIME IN RELATION TO CHANGING CLIMATE AND SURFACE OCEAN CONDITIONS?

Depth:

QUESTION 3:

HAS THE DEPTH RANGE OF MACROURID SPECIES CHANGED OVER TIME WITH CHANGING CLIMATE AND SURFACE-OCEAN CONDITIONS?

QUESTION 4:

HAS THE DEPTH RANGE OF MERLUCCIUS PRODUCTUS CHANGED OVER TIME WITH CHANGING CLIMATE AND SURFACE-OCEAN CONDITIONS?

1.) <u>Macrourid Species Complex' Abundance</u>

Fish Variable			Species Included	Data Source	Climate/Surface Condition Variables					Time-Series
Macrourid species complex' abundance (# ind. /hr/ week)			Coryphaenoides armatus Corphaenoides leptolepis Coryphaenoides yaquinae	Sta. M	POC	C_N	MEI	SST	NPP	2011-2017
									tical Tasta	
Question #	Question Section			Question &	Hypotheses				Statis	tical lests
	A	Has Macrourid species complex' population structure changed over time in relation to changing climate and surface-ocean conditions?								n-Kendall
1		Ho: With increasing changes in climate and surface ocean conditions over time, there is no significant change in Macrourid A species complex' abundance.						Cross Corre	Cross Correlation Analysis	
		<i>Ha:</i> With i	ncreasing changes in climate and	surface ocean condit complex' a	ions over time, bundance.	there is a signi	ficant change in	Macrourid specie	s Spear	man Rank

Cross correlation analysis showed a positive relationship between MEI and Macrourid species complex' abundance (rho = 0.40; p = 0.0000), with the strongest correlation when fish abundance was lagged 56 weeks after MEI data (Refer to Table 1 in Appendix B). There was significant evidence of a moderate monotonic association between these two variables (rho = 0.40; p = 0.0000). An inverse relationship was found between NPP and Macrourid species complex' abundance (rho = -0.35; p = 0.0000), with the strongest correlation present when fish abundance was lagged 50 weeks after NPP data. A significantly weak monotonic trend was exhibited over the time series between these two variables (rho = -0.35; p = 0.0000). An inverse relationship between Macrourid species complex' abundance and C:N was also found (rho = -0.32; p = 0.0000), with the strongest correlation present when fish abundance lagged 67 weeks after changes in C:N data. A significantly weak monotonic relationship between these two variables over time was determined (rho = -0.32; p = 0.0000).

Relating back to the research question (Q1(A)), results from the Mann-Kendall tests, carried out on Macrourid species complex' abundance, showed a significant increasing trend in the series over time (alpha < 0.05; See Table # above). Results from the Mann-Kendall analysis, in combination with the findings mentioned above from CC and Spearman rank analyses, allow us to reject the null hypothesis (Ho) and accept the alternative hypothesis (Ha), which is stated below.

(Q1(A)): HAS MACROURID SPECIES COMPLEX' POPULATION STRUCTURE CHANGED OVER TIME IN RELATION TO CHANGING CLIMATE AND SURFACE-OCEAN CONDITIONS?

HA: WITH INCREASING CHANGES IN CLIMATE AND SURFACE OCEAN CONDITIONS OVER TIME, SPECIFICALLY IN REGARD TO MEI, NPP AND C:N, THERE IS A SIGNIFICANT INCREASE IN MACROURID SPECIES COMPLEX' ABUNDANCE; THEREFORE, THERE IS A SIGNIFICANT CHANGE IN MACROURID SPECIES COMPLEX' POPULATION STRUCTURE OVER TIME.

Question #	Question Section		Statis	tical Tests						
1		Has M	acrourid sp.' population structure	changed over time in	n relation to cho	anging climate o	and surface-oce	an conditions?	Man	n-Kendall
	D	Ho: Wi	th increasing changes in climate a	nd surface ocean co species' bo	nditions over tir dy lengths.	me, there is no	significant chan	ge in Macrourid	Cross Corr	elation Analysis
		Ha: W	ith increasing changes in climate a	and surface ocean co species' bo	nditions over ti dy lengths.	me, there is a s	ignificant chang	e in Macrourid	Spear	man Rank
-	Fich Variable Species Included Data Source Climate/Surface Condition Variables								Time-Series	
Macrourid length (cm) (Regional)		Nezumia stelgidolepis Albatrossia pectoralis Coryphaenoides acrolepis Nezumia liolepis	Mined	POC	C_N	MEI	SST	NPP	2011-2017	

2.) <u>Macrourid Body Length</u>

The two strongest negative relationships associated with Macrourid body length were with NPP and C:N (NPP: rho = -0.58, p = 0.0000; C:N: rho = -0.53, p = 0.0004). Strongest correlations were found when fish lengths were lagged 56 weeks after NPP data and 39 weeks after C:N data (Refer to Fig. in Appendix). There is significant evidence of independent moderate monotonic associations between these two variables and Macrourid length (NPP: rho = -0.58, p = 0.0000; C:N: rho = -0.53, p = 0.0004). A positive relationship between Macrourid length and SST was found (rho = 0.40; p = 0.0005) with the strongest correlation when fish abundance was lagged 30 weeks after SST data. A significantly moderate monotonic relationship existed between these two variables over time (rho = 0.40; p = 0.0005). Lastly, a positive association was found between Macrourid length and MEI (rho = 0.35; p = 0.0023), when fish lengths were lagged 23 weeks behind NPP data. The two variables exhibited a significantly weak monotonic relationship over time (rho = 0.35; p = 0.0023).

Relating back to the research question (Q1(F)), results from the Mann-Kendall tests, carried out on Macrourid lengths, showed a significant increasing trend in the time-

series (alpha < 0.05; See Table # above). Results from the Mann-Kendall analysis, in combination with the findings mentioned above from CC and Spearman rank analyses, allow us to reject the null hypothesis (Ho) and accept the alternative hypothesis (Ha), which is stated below.

(Q1(F)): HAS MACROURID SP.' POPULATION STRUCTURE CHANGED OVER TIME IN RELATION TO CHANGING CLIMATE AND SURFACE-OCEAN CONDITIONS?

HA: WITH INCREASING CHANGES IN CLIMATE AND SURFACE OCEAN CONDITIONS OVER TIME, SPECIFICALLY IN REGARD TO NPP, C:N, SST, AND MEI, THERE IS A SIGNIFICANT INCREASE IN MACROURID LENGTHS; THEREFORE, THERE IS A SIGNIFICANT CHANGE IN MACROURID SP.' POPULATION STRUCTURE OVER TIME.

3.) Local Merluccius productus Biomass

Question #	Question Section		Statis	Statistical Tests						
		Has Merlu	ccius productus' population struct	ure changed over tim	ne in relation to	changing clima	te and surface-o	ocean conditions	? Man	n-Kendall
2	A	Ho: Wit	h increasing changes in climate a	nd surface ocean con productus	ditions over tim biomass.	ie, there is no s	ignificant chang	e in <i>Merluccius</i>	Cross Corre	elation Analysis
		Ha: With increasing changes in climate and surface ocean conditions over time, there is a significant change in Merluccius productus' biomass.							Spear	Spearman Rank
	Fish Veriables		Consister Included	Data Caurea		Climate /Cu	face Canditia			Times Carias
	Fish variables		Species included	Data Source		Climate/Su	race conditio	on variables		Time-Series
<i>M. productus</i> biomass (kg per ha/der) (Regional & Local)		Merluccius productus	Mined	POC	C_N	MEI	SST	NPP	2011-2017	

Overall, local *M. productus*' biomass data were found to have the strongest relationships between the climate/surface ocean condition variables out of the three fish variables exhibiting trends over time. More specifically, SST exhibited the strongest association with local biomass. The relationship between local *M. productus*' biomass (kg per ha/der) data and SST was positive (rho = 0.75; p = 0.0005), with the strongest correlation present when local *M. productus*' biomass was lagged 6 weeks after SST data (Refer to Fig in Appendix). There is significant evidence of a strong monotonic association between these two variables (rho = 0.75; p = 0.0005). A positive relationship between local *M. productus*' biomass and MEI was also found (rho = 0.76; p = 0.0003). The strongest correlation occurred when local *M. productus*' biomass was lagged 15

weeks after MEI data. A significantly strong monotonic relationship was determined between these two variables over time (rho = 0.76; p = 0.0003). Lastly, NPP was found to be inversely associated with local *M. productus*' biomass data (rho = -0.68; p = 0.0027), with the strongest correlation when local *M. productus*' biomass was lagged 2 weeks after NPP data. A significantly strong monotonic relationship existed between these two variables over time (rho = -0.68; p = 0.0027).

Relating back to the research question (Q1(D)), results from the Mann-Kendall tests, carried out on local *M. productus*' biomass, showed a significant increasing trend in the series over time (alpha < 0.05; See Table # above). Results from the Mann-Kendall analysis, in combination with the findings mentioned above from CC and Spearman rank analyses, allow us to reject the null hypothesis (Ho) and accept the alternative hypothesis (Ha), which is stated below.

(Q1(D)): HAS MERLUCCIUS PRODUCTUS' POPULATION STRUCTURE CHANGED OVER TIME IN RELATION TO CHANGING CLIMATE AND SURFACE-OCEAN CONDITIONS?

HA: WITH INCREASING CHANGES IN CLIMATE AND SURFACE OCEAN CONDITIONS OVER TIME, SPECIFICALLY IN REGARD TO SST, MEI, AND NPP, THERE IS A SIGNIFICANT INCREASE IN LOCAL *MERLUCCIUS PRODUCTUS*' BIOMASS; THEREFORE, THERE IS A SIGNIFICANT CHANGE IN *MERLUCCIUS PRODUCTUS*' POPULATION STRUCTURE OVER TIME.

Question #	Question Section	Question & Hypotheses								tical Tests	
		Has Merlu	iccius productus' population struct	ture changed over tin	ne in relation to	changing clima	te and surface o	ocean conditions	? Man	n-Kendall	
2	В	2 B Ho: With increasing changes in climate and surface ocean conditions over time, there is no significant change in Merluccius productus' CPUE.								Cross Corre	elation Analysis
		Ha: With increasing changes in climate and surface ocean conditions over time, there is a significant change in Merluccius productus' CPUE.							Spear	Spearman Rank	
-					1					1	
	Fish Variables		Species Included	Data Source		Climate/Su	rface Conditio	on Variables		Time-Series	
<i>M. productus</i> CPUE (num per ha/der) (Regional & Local)			Merluccius productus	Mined	POC	C_N	MEI	SST	NPP	2011-2017	

4.) Other Noteworthy Relationships

Question #	Question Section			Question &	Hypotheses				Statis	tical Tests	
		Has Merluccius productus' population structure changed over time in relation to changing climate and surface ocean conditions?								Mann-Kendall	
2	2 D Ho: With increasing changes in climate and surface ocean conditions over time, there is no significant change in Merluccius productus' larval abundance.									elation Analysis	
		Ha: With increasing changes in climate and surface ocean conditions over time, there is a significant change in Merluccius productus' larval abundance.								Spearman Rank	
		-							-		
	Fish Variables		Species Included	Data Source	Climate/Surface Condition Variables					Time-Series	
<i>M. productus</i> larval abundance (10 m ²) (Regional)			Merluccius productus	Mined	POC	C_N	MEI	SST	NPP	2011-2017	

Question #	Question Section		Statis	tical Tests							
4		Has i	the depth range of Merluccius prod	ductus changed over	time with chan	ging climate and	l surface ocean	conditions?	Man	n-Kendall	
	 Ho: With increasing changes in climate and surface ocean conditions over time, there is no significant change in Merluccius A A productus' depth distribution. 							Cross Corre	Cross Correlation Analysis		
		Ha: With increasing changes in climate and surface ocean conditions over time, there is a significant change in Merluccius productus' depth distribution.							Spear	Spearman Rank	
	Fish Veriables		Constinue Included	Data Caura		Climate /Su	rfaca Canditic	n Variablas		Time Carles	
	Fish variables		Species included	Data Source		climate/su	race conditio	n variables		Time-Series	
<i>M. productus</i> depth (m) (Regional & Local)		(m))	Merluccius productus	Mined	POC	C_N	MEI	SST	NPP	2011-2017	

The associations discussed below are statistically significant and noteworthy, despite lacking trends over time, as determined by the Mann-Kendall tests. The strongest statistical relationship existed between local *M. productus* depth data and C:N. The relationship between local *M. productus* depth and C:N was positive (rho = 0.90; p = 0.0046), with the strongest correlation when local M. productus depth data were lagged 3 weeks after C:N data. There was a very strong monotonic association between these two variables (rho = 0.90; p = 0.0046). The second strongest relationship existed between local M. productus catch per unit effort (CPUE) data and SST. The relationship between these two variables was positive (rho = 0.86; p < $2.2 e^{-16}$), with the strongest correlation when local *M. productus* CPUE was lagged 6 weeks behind SST data. A very strong statistically significant monotonic association between these two variables over time was found (rho = 0.86; p < $2.2 e^{(-16)}$). The third strongest relationship existed between the same local *M. productus* CPUE data and NPP. The relationship between these two variables was negative (rho = -0.80; p = 0.0001), with the strongest correlation when local *M. productus* CPUE was lagged 2 weeks behind NPP data. A very strong monotonic association between these two variables was determined (rho = -0.80; p =

0.0001). Lastly, *M. productus*' larval abundance was significantly associated with C:N, SST, and NPP. The separate relationships of C:N and SST with *M. productus*' larvae were both positive (C:N: rho = 0.58, p = 0.0000; SST: rho = 0.55, p = 0.0000). The strongest correlation was present when *M. productus*' larval abundance data were lagged 4 weeks after C:N data and when lagged 14 weeks after SST data, respectively. *M. productus*' larvae were determined to have a significantly inverse association with NPP (rho = -0.45; p = 0.0003), with the strongest correlation when *M. productus*' larvae were lagged 2 weeks behind NPP. Results determined independent and statistically significant moderate monotonic associations between *M. productus*' larvae and these three variables (C:N: rho = 0.58, p = 0.0000; SST: rho = 0.55, p = 0.0000; NPP: rho = -0.45; p = 0.0003).

Relating back to the research question (Q1(E)), results from the Mann-Kendall tests, carried out on local *M. productus* CPUE, showed a computed p-value greater than the significance level (alpha = 0.05); and can therefore be interpreted as local *M. productus* CPUE lacking a significant trend in the series over time (See Table # above). Results from the Mann-Kendall analysis, in combination with the findings mentioned above from CC and Spearman rank analyses, lead us to the conclusion that one cannot reject the null hypothesis (Ho), which is stated below.

(Q1(E)): HAS MERLUCCIUS PRODUCTUS' POPULATION STRUCTURE CHANGED OVER TIME IN RELATION TO CHANGING CLIMATE AND SURFACE-OCEAN CONDITIONS?

Ho: WITH INCREASING CHANGES IN CLIMATE AND SURFACE OCEAN CONDITIONS OVER TIME, SPECIFICALLY IN REGARD TO SST AND NPP, THERE IS NO SIGNIFICANT CHANGE IN LOCAL *MERLUCCIUS PRODUCTUS*' CPUE; THEREFORE, THERE IS NO SIGNIFICANT CHANGE OVER TIME IN *MERLUCCIUS PRODUCTUS*' POPULATION STRUCTURE IN RELATION TO LOCAL *MERLUCCIUS PRODUCTUS*' CPUE, SST AND NPP.

The null hypothesis could not be rejected on account of there being no significant trend in local *M. productus* CPUE over time. It is worth noting that local *M. productus* CPUE, and therefore local *M. productus* population structure, did exhibit changes in relation to environmental variables (SST and NPP) despite the lack of a significant trend over time. Similarly, results from the Mann- Kendall test carried out on regional *M*.

productus 'larvae (Q1(I)) showed a computed p-value greater than the significance level (alpha = 0.05); and can therefore be interpreted as regional *M. productus*' larvae lacking a significant trend in the series over time (See Table # above). Results from the statistical analyses lead us to the conclusion that one cannot reject the null hypothesis (Ho), which is stated below.

(Q1(I)): HAS MERLUCCIUS PRODUCTUS' POPULATION STRUCTURE CHANGED OVER TIME IN RELATION TO CHANGING CLIMATE AND SURFACE-OCEAN CONDITIONS?

Ho: WITH INCREASING CHANGES IN CLIMATE AND SURFACE OCEAN CONDITIONS OVER TIME, SPECIFICALLY IN REGARD TO C:N, SST AND NPP, THERE IS NO SIGNIFICANT CHANGE IN REGIONAL *MERLUCCIUS PRODUCTUS*' LARVAE; THEREFORE, THERE IS NO SIGNIFICANT CHANGE OVER TIME IN *MERLUCCIUS PRODUCTUS*' POPULATION STRUCTURE IN RELATION TO REGIONAL *MERLUCCIUS PRODUCTUS* LARVAE, C:N, SST, AND NPP.

The null hypothesis could not be rejected on account of there was no significant trend in regional *M. productus*' larvae over time. It is worth noting that regional *M. productus*' larvae, and therefore *M. productus*' population structure, did exhibit changes in relation to environmental variables (C:N, SST, and NPP) despite the lack of a significant trend over time. Finally, results from the Mann-Kendall test carried out on local *M. productus* depth (Q2(B)) showed a computed p-value greater than the significance level (alpha = 0.05); and can therefore be interpreted as local *M. productus* depth lacking a significant trend in the series over time (See Table # above). Results from the statistical analyses lead us to the conclusion that one cannot reject the null hypothesis (Ho), which is stated below.

(Q2(B)): HAS THE DEPTH RANGE OF MERLUCCIUS PRODUCTUS CHANGED OVER TIME WITH CHANGING CLIMATE AND SURFACE-OCEAN CONDITIONS?

Ho: WITH INCREASING CHANGES IN CLIMATE AND SURFACE OCEAN CONDITIONS OVER TIME, SPECIFICALLY IN REGARD TO C:N, THERE IS NO SIGNIFICANT CHANGE IN LOCAL *Merluccius productus*' depth distribution over time.

The null hypothesis could not be rejected on account of there being no significant trend in local *M. productus* depth over time. Like the two previous variables mentioned above, it is worth noting that local *M. productus* depth, and therefore local *M. productus*'
depth distribution, did exhibit changes in relation to environmental variables (C:N) despite the lack of a significant trend over time. In summary, the answers to the two research questions, determined from the results, are as follows:

Q1 & Q2: With increasing changes in climate and surface-ocean conditions over time, there is a significant change in both Macrourid species' population structure and *M*. *productus* population structure over time.

Q4: With increasing changes in climate and surface-ocean conditions over time, there is no significant change in *M. productus*' depth distribution over time.

Q3: Regarding potential changes in Macrourid species' depth distributions, this question couldn't be answered due to insufficient sampling sizes in the mined Macrourid species depth data.

CHAPTER 4

DISCUSSION & CONCLUSIONS

Findings from this study provide a clearer picture of how marine ecosystems are being affected by changes in climate and surface ocean conditions over time. For instance, shifts in nekton abundance within the California Current ecosystem (CCE) have been observed in various studies (Zeidberg & Robison, 2007; Brodeur et al., 2006; Drazen et al., 2012). Changes in Macrourid abundance over time are driven by carrion fluxes (Drazen et al., 2008; Jeffreys et al., 2010; Drazen et al., 2012). Merluccius productus nekton being the main source of carrion supply for benthopelagic Macrourids (Drazen et al., 2009; Dufault et al., 2009; Drazen et al., 2012). Prior to this study, a similar time-series analysis was conducted for an earlier period, 1989-2007, by Drazen et al. (2012) with the goal of evaluating whether changes in Macrourid abundance over time were driven by changes in *M. productus* carrion supply. The goal of this thesis study, however, was to determine: 1) whether Macrourid spp. and/or M. productus population structures have changed over time in relation to changing climate and surface-ocean conditions and 2) whether the depth ranges of Macrourid spp. and/or M. productus have changed over time with changing climate and surface-ocean conditions. While the two studies sought to answer different questions, results from this thesis study corroborate Drazen et al., 2012's findings for the CCE area, as well as for the vicinity of Sta. M. Comparing observations from the two studies, one presenting weekly averages and one presenting x-scale data, can give broader scale insights into how changes in climate and surface ocean conditions might be indirectly influencing population structures and depth distributions of fish species', expanding our understanding in ways that would not be possible using findings from this study alone. More specifically, comparing findings can increase our understanding of how potential impacts of changing climate and surfaceocean conditions on *M. productus*' population structure and depth distribution could indirectly affect Macrourid populations, due to their close relationship with M. productus. It is worth noting that, although M. productus was a focal species for both studies, fish variables (M. productus biomass & larval abundance vs. M. productus egg abundance) used in the analyses were not directly comparable to one another due to lifestage differences. Despite this, it is still beneficial to compare the findings indirectly, giving insight into trends occurring over the life-history of *M. productus*.

4.1) Macrourid & Merluccius productus Observations (2011-2017)

Comparable to Drazen et al. (2012) findings for the years 1989-2007, we observed continued increases in both local Macrourid abundance at Sta. M and regional Macrourid length over the period 2011-2017. Local Macrourid abundance was moderately associated with MEI (+) and weakly associated with NPP (-) and C:N (-). Local Macrourid body length was moderately associated with NPP (-), C:N (-), and SST (+), and weakly associated with MEI (+). When *M. productus* data from the region were localized within 100 nmi of Sta. M, we observed an increase in *M. productus* biomass over time, despite lacking a regional trend. Additionally, out of the fish variables exhibiting trends over time, local *M. productus* biomass was found to have the strongest relationships with climate/surface condition variables. More specifically, strong associations were found between local *M. productus* biomass and SST (+), MEI (+), and NPP (-). Similarly, local *M. productus* CPUE was also strongly associated with SST (+) and NPP (-). Despite lacking a trend over time, regional *M. productus* larval abundance was also significantly correlated to SST (+) and NPP (-), as well as C:N (+). It is worth noting that *M. productus* larval abundance data were not localized for this study.

Similar to this study, Drazen et al. (2012) observed a significant increase in local Macrourid abundances over the time period 1989-2007 and found significant correlations (+) between *M. productus* egg abundance and local Macrourid abundance at all spatial scales. Aside from a four year gap in data between the two time-series studies, combining findings sheds light on the temporal trend occurring from 1989-2017 at Sta. M, in reguards to increased Macrourid abundances .Increases in *M. productus* egg abundance were followed by increases in Macrourid abundance following a lag of 6-7 months at the 100 nautical miles (nmi) scale (Drazen et al., 2012). Correlations between *M. productus* egg abundances were localized to Sta. M at this smaller spatial scale resolution of 100 nmi, compared to the broader regional (CA) scale. When comparing Macrourid abundances at Sta. M with

annual estimates of *M. productus* female spawning stock biomass, for the entire North American west coast, no significant correlations were found. Drazen et al. (2012) also observed a significant increase in average local Macrourid size over time. Combining Macrourid size and abundance allowed for biomass estimates to be assessed, in which a \sim 6 fold increase in Macrourid biomass was observed over the time period. Changes in local Macrourid population structure was determined to follow metrics for *M. productus* biomass, suggesting *M. productus* carrion supply may be the driving factor contributing to changes in some Macrourid population structures (Drazen et al., 2012).

As previously mentioned, results from this study found increases in local M. productus biomass, local Macrourid abundance, and regional Macrourid length over time. Despite lacking Macrourid biomass data, observed increases in the abundance and length of local Macrourids over time matches findings from Drazen et al. (2012). The two studies used different analyses to assess Macrourid and *M. productus* populations, however, we can still compare the findings indirectly to gain further insight into what the separate results might signify. Observed positive correlations between local M. productus egg abundance and local Macrourid abundance (Drazen et al., 2012) suggest that an increase in one variable would lead to an increase in the other. M. productus larval abundance was not found to increase over time in this thesis study, however, an increasing trend over time may have been observed if the larval data had been localized relative to Sta. M. Regardless, Drazen et al. (2012) found that changes in local M. productus biomass were followed by changes in local Macrourid population structure. Findings from this thesis study compliment this finding, in that we observed increases in both local M. productus biomass and local Macrourid abundance and length. It is also worth noting that all local *M. productus* variables (biomass, CPUE, & larval abundance) were all significantly correlated with SST (+) and NPP (-), suggesting changes in local larval abundance, driven by SST and/or NPP, are likely to somewhat resemble changes in local adult biomass.

4.2) Spatial Scale & Location Significance

As previously stated, *M. productus* carrion has been found to be the largest and most important food source for Macrourids (Drazen et al., 2009; Dufault et al., 2009), making up 61% of pelagic biomass in the California Current ecosystem (CCE) (Agostini et al., 2006; Drazen et al., 2012). For this study, spatial scale was a determining factor in the significance of the observed findings. When localized to Sta. M, we observed significant increases in Macrourid abundance and length, as well as increased *M. productus* biomass over time. When cross correlations were run for the broader region of California waters, these trends were not significant. Similar to findings from this study, Drazen et al., 2012 found that the relationships between *M. productus* egg abundance and Macrourid abundance increased significantly when localized to Sta. M at 100 nmi. Sta. M is unique for many reasons, including spatial, temporal and biogeochemical processes (Smith et al., 2006); which may explain the significance found when localizing the datasets. In regard to Macrourid population structure dynamics, various studies suggest location of *M. productus* spawning may be more of a factor than spawning stock size (Bailey 1982; Mullin et al., 2000; Horn & Smith, 1997; Drazen et al., 2012).

Station M, underlying part of the California Current (CC) (Smith et al., 2017), lies within the inshore region known as the Southern California Bight (SCB) (Bograd et al., 2015). POC flux varies seasonally and interannually at this site (Ruhl, 2008). The CCE is an important and highly productive upwelling region due to Ekman transport (Pickett & Paduan, 2003; Bograd et al., 2019). The California Undercurrent (CU) has a strong impact on the SCB, particularly in regard to areas over the continental slope (Bograd et al., 2015; Bograd et al., 2019). The southern CCE region acts as a transitional boundary between various water masses, including the Pacific Equatorial Water (PEW), Eastern North Pacific Central Water (ENPCW), and the Pacific Subarctic Upper Water (PSUW) (Brodeur et al., 2003; Bograd et al., 2019). PEW, ENPCW, and PSUW strongly influence the CU, lateral shifts in the North Pacific subtropical gyre boundary, and the CC respectively (Bograd et al., 2019). These water masses have different physical properties and their separate relative influences can vary interannually (Bograd et al., 2019). Such variations can have multi-scale consequences on ecosystem structure and biogeochemical processes (Bograd et al., 2015).

Merluccius productus is a highly valued target species for fisheries around the world, including US and Canadian fisheries (Agostini et al., 2006). Fish yields between these two countries are strongly influenced by interannual biomass distribution of the species. Substantial variability in year-class strength upsets this distribution, potentially impacting the fisheries in a negative way (Agostini et al., 2006). Studies found that yearclass strength may depend on survival of the larval stages, specifically the late stage (Hollowed, 1992; Smith, 1985b). Locations such as Sta. M and the SCB are significant for the US and Canadian fisheries for this reason. Studies have determined the SCB, where *M. productus* nursery grounds are located, and regions west of Point Conception (Sta. M) are favorable to the survival of *M. productus* larvae, and late stage larvae have been shown to congregate in these areas regularly (Hollowed, 1992; Walker et al., 1987). This is largely due to the layered physical structure of this particular region (Sta. M and SCB) (Hollowed, 1992). The survival of *M. productus* larvae may rely on reduced offshore advection (Bailey et al., 1982; Bailey, 1981). The layered structure of the region results in subsurface waters from the CU being directed onshore while surface waters stemming from the CC are directed offshore (Hollowed, 1992). Nearshore, offshore transport is minimized (Morgan & Fisher, 2010). This, combined with the diurnal vertical migration of *M. productus* larvae, reduces offshore advection, allowing larvae to remain in the study region where prey abundance is high (Hollowed, 1992).

Prey availability is another important factor contributing to the survival or mortality of *M. productus* larvae (Bailey, 1981). Prey is more than abundant in the study area (Sta. M) due to the high productivity of the region. Extended periods of high-nutrient conditions caused by upwelling in this region result in high productivity (Pickett & Paduan, 2003; Bograd et al., 2019). Increases in phytoplankton from optimal-nutrient conditions results in high zooplankton abundances (Danger et al., 2007; Vanni, 1987). *Calanus pacificus* is a primary prey item for *M. productus* adults. Increased abundances of phytoplankton cause increased production in this prey species, which are then prey for larval *M. productus* (Hollowed, 1992). This is in line with findings from this thesis study, in that NPP was found to be significantly and positively associated with *M. productus* regional biomass and CPUE. Local *M. productus* biomass and CPUE were found exhibiting inverse trends associated with NPP. This makes sense in theory because *M. productus* adults migrate north, a long distance away from Sta. M, to feed and are less dependent on NPP associated with the Sta. M region. The influence of NPP on *M. productus* larval abundance, Macrourid length, and local Macrourid abundance in this study were less clear. Significant inverse associations between NPP and these three variables were observed. It is possible that abundance changes of dominant zooplankton species could result in a loss of prey for *M. productus* larvae, negatively affecting Macrourid population structure as well (Brodeur et al., 2006). Another possibility is that observed correlation directions associated with the relationships between NPP and Macrourid length/*M. productus* larvae could potentially switch to positive correlations if length and larval abundance data were localized to Sta. M.

Carbon to nitrogen ratios (C:N) were selected as a food supply variable focused on in this study and goes hand in hand with NPP. Higher C:N ratios in mixed layer waters signifies the use of nitrogen for primary production, resulting in increased food supply for larvae (Körtzinger et al., 2001). Higher C:N ratios also denote more refractory, lower food quality at abyssal depths and hence less nitrogen compared to carbon (Burkhardt et al., 1999). By contrast, lower C:N ratios in mixed layer waters signifies higher food quality in the sinking particulate matter food for Macrourids in the deep sea (Körtzinger et al., 2001). C:N ratios were significantly and positively associated with *M. productus* larval abundance. The influence of C:N ratios on adult fishes in this study were less clear. Significant inverse associations between C:N ratios and local Macrourid abundance and length were observed. These findings make sense in theory, in that higher C:N ratios signify lower quality food for Macrourids in the deep-sea (Ken Smith, pers. comm.), and increased food supply for larvae in the shallower mixed layer waters. Not surprisingly, C:N had the strongest correlations with *M. productus* (+) and Macrourid depth ranges (-) among the environmental variables analyzed in this study.

In addition to spawning location, the spawning migration route of *M. productus* may be another potential factor contributing to the significance of the localized small

spatial scale in relation to Station M and fish associations. Factors including temperature, depth, upwelling, and currents within the CCE contribute significantly to the migration route taken by *M. productus* adult fishes (Peterson et al., 2006; Bailey et al., 1982). Upwelling plays a significant role in the life-stages of *M. productus*, in that for at least half of the year, adults reside in upwelling regions (Cushing, 1971). M. productus travel south at a depth of 350-400 m via the CC to warmer offshore waters in the fall and winter, in order to spawn where larvae can take advantage of productive upwelling (Cushing, 1971; Drazen et al., 2012). Point Conception is considered the southern boundary of the major CCE upwelling zone (Cudaback et al., 2005). The location of spawning can vary slightly with temperature, with observed shifts towards the equator during cold years and shifts poleward during warm years (near Point Conception) (Agostini, 2005). Winds north of Point Conception influence inner shelf currents, specifically the inshore Davidson current (Mazzini et al., 2014). Adult fish take advantage of this countercurrent as the upwelling season advances, riding it north in the spring and summer at a depth of <100 m in order to feed (Cushing, 1971; Stauffer, 1985). Studies have suggested changes in Macrourid length and abundance may be due to migration patterns associated with M. productus migrations (Drazen et al., 2012; Priede et al., 2003; King & Priede, 2008), which aligns with results from this study.

As stated above, temperature plays a significant role in the patterns exhibited by *M. productus* life stages throughout the year (Agnostini, 2005). SST has an effect on spawning location (Agnostini, 2005) and NPP (Körtzinger et al., 2001), which in turn affects *M. productus* year-class strength and larval abundance (Hollowed, 1992), consequently affecting local Macrourid population structure (Drazen et al., 2012). Results from this study reflect this, in that local *M. productus* larvae and adult biomass, as well as local Macrourid length, are significantly associated with SST. SST is one of the six variables associated with the MEI index (Sukresno, 2010); an assessment of ENSO that considers atmospheric and oceanic variables of tropical Pacific (Wolter & Timlin, 2011). Studies have shown source water contributions (PEW, ENPCW, and PSUW) of the CCE are affected by the ENSO cycle, specifically in relation to PEW during La Nina events (Bograd et al., 2019). During these events, changes in depth distribution of this water

source, in combination with fluctuations of upwelling strength, result in primary producers gaining better access to the nutrient-rich waters below (Bograd et al., 2019). During such events, scientists have observed southern shifts in *M. productus* spawning activity (Drazen et al., 2012). Findings from this study compliment this, in that MEI was found to be positively associated with local *M. productus* biomass, Macrourid abundance and length.

4.3 Potential Impacts of Climate and Fishing Pressure

Climate and fishing pressure can have drastic impacts on marine ecosystems, including the deep-sea (Overland et al., 2010). Increases in anthropogenic climate change has the potential to dramatically alter biogeochemical processes within the world's oceans, and the population structures of important fish species targeted by fisheries (Doney, 2010). In light of anthropogenic climate change, ENSO events have become more frequent and have increased in intensity in recent years, and this is predicted to continue in the future (Cai et al., 2015). In addition, ENSO events have been assumed to resemble future warming events (Meehl et al., 2000), which are predicted to cause sea level rise and increases in surface water temperatures along the coast, upwelling, and stratification (Hoegh-Guldberg & Bruno, 2005; DiNezio et al., 2010). Additionally, thermocline depth is expected to get deeper (DiNezio et al., 2010). La Nina events have also been found to impact near surface expression of source waters mentioned above, and this is known to cause amplified low pH and hypoxic events (Bograd et al., 2019). Macrourid and *M. productus* populations residing in the CCE could be affected by changes in climate in various ways, including those residing in the vicinity of Point Conception and the SCB. For instance, the location of *M. productus* spawning shifts in regard to temperature, moving poleward in warm years (Agnostini, 2005). As previously stated, *M. productus* is one of the most important and largest commercial groundfish species in the U.S. (Drazen et al., 2012; Zeidberg & Robison, 2007), and shifts in the distribution of biomass can affect the annual yields for Canadian and US fisheries (Agostini et al., 2006). In addition, the most valuable fish migrate further north (Emmett et al., 2006; Agostini et al., 2006). Increased surface temperature may result in further migrational shifts northward, potentially benefitting northern US and Canadian fisheries.

By contrast, these shifts could potentially reduce abundance in waters off California (Emmett et al., 2006). Aside from shifts in distribution, surface warming may also cause declines in zooplankton-copepod abundances (Smith, 1985a) and community composition changes due to nutrient changes (Pinhassi et al., 2006; Suikkanen et al., 2013) and greater offshore transport (Bakun et al., 2015). Declines in zooplanktoncopepod abundances could mean a decrease in the food source for *M. productus* larvae (Sumida & Moser, 1980) (reducing year-class success) and thus a decrease in food for Macrourids and other species relying on *M. productus* carrion (Bailey, 1981; Hollowed, 1992; Dufault et al., 2009; Brodeur et al., 2003). By contrast, surface warming may favor *M. productus* populations in California waters. During warmer years, spawning locations shift poleward near Point Conception, as previously stated. Spawning in this location reduces offshore advection for larvae, increasing their chances of survival. This study's findings are in line with this assumption because we found positive correlations between SST and *M. productus* larval abundance. Reduction in year-class success from shifts in larval distributions may negatively impact fisheries in future years as well (Agostini et al., 2006). Increased temperature also leads to low productivity (Behrenfeld et al., 2006). Many species will likely move closer to the coast and northward (Cushing, 1971), however, some species may not be able to. For instance, the habitat range of the deep-sea Macrourid C. yaquinae is limited to the North Pacific central gyre (Drazen et al., 2012). C. yaquinae may not be able to adapt or relocate as the regions with low productivity expand in the gyre, likely reducing the abundance of their food source (Drazen et al., 2012).

Finally, a big impact of climate change on marine ecosystems (i.e. a significantly higher stress level) stems from declining midwater oxygen concentrations and shoaling of oxygen minimum zones (OMZs) (Booth et al., 2012). OMZ expansion in the Eastern Pacific was found to be correlated with CU intensification (Czeschel et al., 2012). Such intensification of the CU can significantly affect the CCE (Bograd et al., 2019). CU intensification is one of the factors contributing to the poleward shift in *M. productus* feeding and spawning grounds (Bograd et al., 2019). Shoaling of the thermocline and OMZs could potentially affect vertical migrations of *M. productus* larvae and could

compress the vertical habitat of adult fish as well, making them more vulnerable to commercial fishing (Gilly et al., 2013). One study found that in summer 1955, a shortterm shift in water masses (from PSUW to PEW) at certain depths caused variations in "depth and thermosteric anomaly surfaces on either side of the transition" (Blanton & Pattullo, 1970]. These two OMZ zones, were separated by an oxygen enriched zone caused by vertical mixing. The enriched zone was located within my study site area near Point Conception (Blanton & Pattullo, 1970). This may be a clue as to why the location of Sta. M was primarily significant for correlations relating to *M. productus* (i.e. local biomass, local CPUE, and local depth range). This region and regions with similar characteristics (i.e. upwelling and location relative to currents) should be considered in future fishery and climate-mitigation decisions, as they are extremely important areas for fisheries and marine ecosystems. Future climate variation could alter the water properties of source waters (Bograd et al., 2015), causing variation in OMZ zones and other factors. Locations such as Sta. M may provide safe zones in the future for *M. productus* and other species in light of future climate impacts; unless upwelling of hypoxic waters occurs (Booth et al., 2012); in which the impacts would be devastating. Lastly, climate-related factors such as OMZ shoaling has resulted in a range expansion for the Humboldt Squid, potentially contributing to top-down forcing on local M. productus populations (Gilly et al., 2013; Zeidberg & Robison, 2007). Range expansions and increases in Humboldt Squid abundance within these ranges have been correlated with declines in M. productus biomass (Mooney & Zavaleta, 2016; Zeidberg & Robison, 2007), and studies have determined this species' principal prey over the coastal California slope and shelf is M. productus. (Zeidberg & Robison, 2007). Increased predation, schooling structure changes, and competition between and among Humboldt squid and M. productus could potentially threaten fisheries, Macrourid populations, and the ecosystems they reside in (Gilly et al., 2013).

Regarding fisheries, increased bycatch from commercial fishing can positively impact Macrourid populations short term due to increased carrion fluxes (Dayton et al., 2003; Drazen et al., 2012; Wallace, 2015). However, commercial overfishing may lead to a reduction in carrion flux in the deep-sea, negatively impacting Macrourids and other species long-term (Wallace, 2015; Drazen et al., 2012). This is especially true for species whose migration patterns reflect that of *M. productus*, and who may have grown dependent on bycatch as their primary food source (Gage et al., 2005). Macrourids and other scavengers may expand their populations in response to bycatch, potentially altering the balance of ecological function (Gage et al., 2005). Additionally, overfishing can potentially lead to shifts in population structure. Such shifts could negatively impact recruitment of *M. productus* larvae, as well as cause shifts in the location and range of adult spawning (Horn & Smith, 1997). Future studies should seek to understand trends and the state of primary producers and zooplankton populations, as well as *M. productus* and Macrourid stocks, and bycatch distribution ranges/intensities in order to better predict, mitigate, and adapt to climate change and commercial fishing related impacts.

REFERENCES

Agnostini VN. 2005. Climate ecology and productivity of Pacific sardine (Sardinops sagax) and hake (Merluccius productus). PhD diss., University of Washington.

Agostini VN, Francis RC, Hollowed AB, Pierce SD, Wilson C, Hendrix AN. 2006. The relationship between Pacific hake (Merluccius productus) distribution and poleward subsurface flow in the California Current System. Can. J. Fish. Aquat. Sci. 63(12):2648-2659.

Allen LG, Pondella DJ, Horn MH. 2006. The ecology of marine fishes: California and adjacent waters. Berkeley (CA): University of California Press.

Alverson DL, Larkins HA. 1969. Status of knowledge of the Pacific hake resource. Calif. Coop. Ocean. Fish. Invest. Rep. 13:24-31.

Arteaga L, Haëntjens N, Boss E, Johnson KS, Sarmiento JL. 2018. Assessment of export efficiency equations in the Southern Ocean applied to satellite-based net primary production. J. Geophys. Res. Oceans. 123:2.

Bailey DM, Ruhl HA, Smith KL Jr. 2006. Long-term changes in benthopelagic fish abundance in the abyssal N.E. Pacific Ocean. Ecology. 87(3):549-555.

Bailey KM. 1981. Larval transport and recruitment of Pacific hake Merluccius productus. Mar. Ecol. Prog. Ser. 6(1):1-9.

Bailey KM, Francis RC, Stevens PR. 1982. The life history and fishery of Pacific whiting, Merluccius productus. NOAA Rev.

Bakun A, Black BA, Bograd SJ, Garcia-Reyes M, Miller AJ, Rykaczewski RR, Sydeman WJ. 2015. Anticipated effects of climate change on coastal upwelling ecosystems. Curr. Clim. Change Rep. 1(2):85-93.

Baldrighi E, Giovannelli D, D'Errico G, Lavaleye M, Manini E. 2017. Exploring the relationship between macrofaunal biodiversity and ecosystem functioning in the deep sea. Front. Mar. Sci. 4.

Beamish RJ, McFarlane GA, King JR. 2005. Migratory patterns of pelagic fishes and possible linkages between open ocean and coastal ecosystems off the Pacific coast of North America. Deep Sea Res. Part II Top. Stud. Oceanogr. 52(56):739-755.

Behrenfeld MJ, O'Malley RT, Siegel DA, McClain CR, Sarmiento JL, Feldman GC, Milligan AJ, Falkowski PG, Letelier RM, Boss ES. 2006. Climate-driven trends in contemporary ocean productivity. Nature. 444(7120):752-755.

Bellido JM, Santos MB, Pennino MG, Valeiras X, Pierce GJ. 2011. Fishery discards and bycatch: solutions for an ecosystem approach to fisheries management? Hydrobiologia. 670(1):317.

Berner C, Bertos-Fortis M, Pinhassi J, Legrand C. 2018. Response of microbial communities to changing climate conditions during summer cyanobacterial blooms in the Baltic Sea. Front. Microbiol. 9.

Blanton J, Pattullo JG. 1970. The subsurface boundary between subarctic Pacific water and Pacific equatorial water in the transition zone off Southern California. Limnol. Oceanogr. 15:606-614.

Bograd SJ, Schroeder ID, Jacox MG. 2019. A water mass history of the Southern California current system. Geophys. Res. Lett. 46(12):6690-6698.

Bograd SJ, Buil MP, Lorenzo ED, Castro CG, Schroeder ID, Goericke R, Anderson CR, Benitez-Nelson C, Whitney FA. 2015. Changes in source waters to the Southern California Bight. Deep Sea Res. Part II Top. Stud. Oceanogr. 112:42-52.

Booth JAT, McPhee-Shaw EE, Chua P, Kingsley E, Denny M, Phillips R, Bograd SJ, Zeidberg LD, Gilly WF. 2012. Natural intrusions of hypoxic, low pH water into nearshore marine environments on the California coast. Cont. Shelf Res. 45:108-115.

Botsford LW, Moloney CL, Hastings A, Largier JL, Powell TM, Higgins K, Quinn JF. 1994. The influence of spatially and temporally varying oceanographic conditions on meroplanktonic metapopulations. Deep Sea Res. Part II Top. Stud. Oceanogr. 41(1):107-145.

Brander KM. 2007. Global fish production and climate change. Proc. Natl. Acad. Sci. U.S.A. 104(50):19709-19714.

Brierley AS, Kingsford MJ. 2009. Impacts of climate change on marine organisms and ecosystems. Curr. Biol. 19(14):R602-R614.

Brodeur RD, Pearcy WG, Ralston S. 2003. Abundance and distribution patters of nekton and micronekton in the Northern California Current Transition Zone. J. Oceanogr. 59(4):515-535.

Brodeur RD, Ralston S, Emmett RL, Trudel M, Auth TD, Phillips AJ. 2006. Anomalous pelagic nekton abundance, distribution, and apparent recruitment in the northern California Current in 2004 and 2005. Geophys. Res. Lett. 33(22).

Buesseler KO, Trull TW, Steinberg DK, Silver MW, Siegel DA, Saitoh SI, Lamborg CH, Lam PJ, Karl DM, Jiao NZ, Honda MC. 2008. VERTIGO (VERtical Transport In the Global Ocean): A study of particle sources and flux attenuation in the North Pacific. Deep Sea Res. Part II Top. Stud. Oceanogr. 55(14-15):1522-1539.

Burkhardt S, Zondervan I, Riebesell U. 1999. Effect of CO2 concentration on C:N:P ratio in marine phytoplankton: a species comparison. Limnol. Oceanogr. 44:683-690.

Byrne M, Przesławski R. 2013. Multistressor impacts of warming and acidification of the ocean on marine invertebrates' life histories. Integr. Comp. Biol. 53(4):582-596.

Cai W, Santoso A, Wang G, Yeh SW, An SI, Cobb KM, Collins M, Guilyardi E, Jin FF, Kug JS, Lengaigne M. 2015. ENSO and greenhouse warming. Nat. Clim. Change. 5(9):849-859.

Carrera M, Cañas B, Piñeiro C, Vázquez J, Gallardo JM. 2006. Identification of commercial hake and grenadier species by proteomic analysis of the parvalbumin fraction. Proteomics. 6(19):5278-5287.

Chavez FP, Messié M, Pennington JT. 2011. Marine primary production in relation to climate variability and change. Annu. Rev. Mar. Sci. 3: 227-60.

Cheung WWL, Lam VWY, Sarmiento JL, Kearney K, Watson R, Zeller D, Pauly D. 2010. Large-scale redistribution of maximum fisheries catch potential in the global ocean under climate change. Glob. Change Biol. 16(1):24–35.

Clark M. 2001. Are deepwater fisheries sustainable? – the example of orange roughy (Hoplostethus atlanticus) in New Zealand. Fish. Res. 51(2-3):123-135.

Cohen DM, Inada T, Iwamoto T, Scialabba N. 1990. FAO species catalogue. Gadiform fishes of the world (Order Gadiformes). An annotated and illustrated catalogue of cods, hakes, grenadiers, and other gadiform fishes known to date. FAO Fish. Synop. 10(125):442.

Cudaback CN, Washburn L, Dever E. 2005. Subtidal inner-shelf circulation near Point Conception, California. J. Geophys. Res. Oceans. 110(C10).

Cushing DH. 1971. Upwelling and the production of fish. Adv. Mar. Biol. 9:255-334.

Czeschel R, Stramma L, Johnson GC. 2012. Oxygen decreases and variability in the eastern equatorial Pacific. J. Geophys. Res. Oceans. 117(C11).

Danger M, Oumarou C, Benest D, Lacroix G. 2007. Bacteria can control stoichiometry and nutrient limitation of phytoplankton. Funct. Ecol. 21(2):202-210.

Danovaro R, Gambi C, Dell'Anno A, Corinaldesi C, Fraschetti S, Vanreusel A, Vincx M, Gooday AJ. 2008. Exponential decline of deep-sea ecosystem functioning linked to benthic biodiversity loss. Curr. Biol. 18(1):1-8.

Davies AJ, Roberts JM, Hall-Spencer J. 2007. Preserving deep-sea natural heritage: emerging issues in offshore conservation and management. Biol. Cons. 138(3-4):299-312.

Dayton PK, Thrush S, Coleman FC. 2003. Ecological effects of fishing. Pew Oceans Comm. Rep. Arlington (VA) (USA).

DiNezio P, Clement A, Vecchi G. 2010. Reconciling differing views of tropical Pacific climate change. Eos, Trans. AGU. 91(16):141-142.

Doney SC. 2010. The growing human footprint on coastal and open-ocean biogeochemistry. Science. 328(5985):1512-1516.

Doney SC, Ruckelshaus M, Duffy JE, Barry JP, Chan F, English CA, Galindo HM, Grebmeier JM, Hollowed AB, Knowlton N, Polovina J. 2011. Climate change impacts on marine ecosystems. Annu. Rev. Mar. Sci.

Doney SC, Ruckelshaus M, Emmett DJ, Barry JP, Chan F, English CA, Galindo HM, Grebmeier JM, Hollowed AB, Knowlton N, Polovina J, Rabalais NN, Sydeman WJ, Talley LD. 2012. Climate change impacts on marine ecosystems. Annu. Rev. Mar. Sci. 4(1):11–37.

Drazen JC, Bailey DM, Ruhl HA, Smith KL. 2012. The role of carrion supply in the abundance of deep-water fish off California. PLoS One. 7(11).

Drazen JC, Popp BN, Choy CA, Clemente T, Forest LD, Smith KL Jr. 2008. Bypassing the abyssal benthic food web: Macrourid diet in the eastern North Pacific inferred from stomach content and stable isotopes analyses. Limnol. Oceanogr. 53(6):2644-2654.

Drazen JC, Phleger CF, Guest MA, Nichols PD. 2009. Lipid composition and diet inferences in abyssal macrourids of the eastern North Pacific. Mar. Ecol. Prog. Ser. 387: 1-14.

Dufault, AM, Marshall K, Kaplan IC. 2009. A synthesis of diets and trophic overlap of marine species in the California Current. U.S. Dept. Commer., NOAA Tech. Memo. NMFS-NWFSC-103:81.

Emmett RL, Krutzikowsky GK, Bentley P. 2006. Abundance and distribution of pelagic piscivorous fishes in the Columbia River plume during spring/early summer 1998–2003: Relationship to oceanographic conditions, forage fishes, and juvenile salmonids. Prog. Oceanogr. 68:1-26.

Endo H, Okamura O. 1992. New records of the abyssal grenadiers Coryphaenoides armatus and C. yaquinae from the western North Pacific. Jpn. J. Ichthyol. 38(4):433-437.

Essington T, Klinger T, Conway-Cranos T, Buchanan J, James A, Kershner J, Logan I, West J. Bentho-pelagic fish in Puget Sound. In: 2011. Puget Sound Science Update. Apr. 2011 version. Tacoma (WA): Puget Sound Partnership.

Fiksen O, Jørgensen C. 2011. Model of optimal behavior in fish larvae predicts that food availability determines survival, but not growth. Mar. Ecol. Prog. Ser. 432:207–219.

Fisheries and Oceans Canada. 2018. Pacific Hake/Pacific Whiting. Canada [Internet]. [cited 2018 Oct. 18]. Available from: http://www.dfo-mpo.gc.ca/species-especes/profilesprofils/pacifichake-whiting-merlupacifique-nord-eng.html

Fraschetti S, Terlizzi A, Benedetti-Cecchi L. 2005. Patterns of distribution of marine assemblages from rocky shores: Evidence of relevant scales of variation. Mar. Ecol. Prog. Ser. 296:13–29.

Gage JD, Tyler PA. 1992. Deep-sea biology: a natural history of organisms at the deepsea floor. Cambridge, UK: Cambridge University Press. (No. 504.42 GAG).

Gage JD, Roberts JM, Hartley JP, Humphery JD. 2005. Potential impacts of deep-sea trawling on the benthic ecosystem along the northern european continental margin: A review. Am. Fish. Soc. Symp. 41:503-517.

Gilly WF, Beman JM, Litvin SY, Robison BH. 2013. Oceanographic and biological effects of shoaling of the oxygen minimum zone. Annu. Rev. Mar. Sci. 5:393-420.

Gjerde KM. 2006. Ecosystems and biodiversity in deep waters and high seas. UNEP/Earthprint. (178).

Hamilton LC, Haedrich RL, Duncan CM. 2004. Above and below the water: social/ecological transformation in northwest Newfoundland. Popul. Environ. 25(6):195-215.

Heath MR. 1992. Field investigations of the early life stages of marine fish. Adv. Mar. Biol. 28(C):1–174.

Hinrichsen HH. 2009. Biological processes and links to the physics. Deep Sea Res. Part II Top. Stud. Oceanogr. 56(21–22):1968–1983.

Hiroaki S, Keriko JM, Yoshida K. 2002. Utilization of the lipids of the deep-sea grenadier: the high content of ceramide derivatives in the internal organ lipids of genus Coryphaenoides. Fish. Sci. 68(sup2):1426-1429.

Hoff GR, Stevenson DE, Orr JW. 2015. Guide to the gadiform fishes of the eastern North Pacific. U.S. Dep. Commer., NOAA Tech. Memo. [Internet]. [cited 2019 Oct. 20]; NMFS-AFSC-309:68. Available from: https://www.afsc.noaa.gov/Publications/AFSC-TM/NOAA-TM-AFSC-309.pdf

Hoegh-Guldberg O, Bruno JF. 2005. The impact of climate change on the world's marine ecosystems. Science. 308:541.

Hollowed AB. 1992. Spatial and temporal distributions of Pacific hake, Merluccius productus, larvae and estimates of survival during early life stages. Calif. Coop. Ocean. Fish. Invest. Rep. 33:100-123.

Horne JK, Smith PE. 1997. Space and time scales in Pacific hake recruitment processes: latitudinal variation over annual scales. Calif. Coop. Ocean.Fish. Invest. Rep. 90-102.

Iwamoto T. 1999. Order Gadiformes Macrouridae. Calif. Acad. Sci., USA: 977.

Iwamoto T. 2015. Coryphaenoides leptolepis. The IUCN Red List of Threatened Species. [Internet]. [cited 2019 Nov 18]. Available from: http://dx.doi.org/10.2305/IUCN.UK.2015-4.RLTS.T15522140A15603525.en

Iwamoto T, Nakayama N, Shao KT, Ho HC. 2015. Synopsis of the grenadier fishes (Gadiformes; Teleostei) of Taiwan. Proc. Calif. Acad. Sci. 62:31-126.

Jamieson AJ, Priede IG, Craig J. 2012. Distinguishing between the abyssal macrourids *Coryphaenoides yaquinae* and *C. armatus* from in situ photography. Deep Sea Res. Part I Oceanogr. Res. Pap. 64:78-85.

Jeffreys RM, Lavaleye MSS, Bergman MJN, Duineveld GCA, Witbaard R, Linley T. 2010. Deep-sea macrourid fishes scavenge on plant material: evidence from in situ observations. Deep Sea Res. Part I Oceanogr. Res. Pap. 57(4):621-627.

Johnson CR, Banks SC, Barrett NS, Cazassus F, Dunstan PK, Edgar GJ, Frusher SD, Gardner C, Haddon M, Helidoniotis F, and Hill KL. 2011. Climate change cascades: Shifts in oceanography, species' ranges and subtidal marine community dynamics in eastern Tasmania. J. Exp. Mar. Biol. Ecol. 400(1-2):17-32.

Kamykowski D, Zentara SJ, Morrison JM, Switzer AC. 2002. Dynamic global patterns of nitrate, phosphate, silicate, and iron availability and phytoplankton community composition from remote sensing data. Glob. Biogeochem. Cycles. 16(4):25-1-25–29.

Karnosky DF. 2003. Impacts of elevated atmospheric CO2 on forest trees and forest ecosystems: knowledge gaps. Environ. Int. 29(2-3):161-169.

Katz T, Yahel G, Yahel R, Tunnicliffe V, Herut B, Snelgrove P, Crusius J, Lazar B. 2009. Groundfish overfishing, diatom decline, and the marine silica cycle: Lessons from Saanich Inlet, Canada, and the Baltic Sea cod crash. Glob. Biogeochem. Cycles. 23:4.

Katz T, Yahel G, Tunnicliffe V, Herut B, Whitney F, Snelgrove PV, Lazar B. 2016. The silica cycle in a Northeast Pacific fjord; the role of biological resuspension. Prog. Oceanogr. 147:10-21.

King JR, Agostini VN, Harvey CJ, McFarlane GA, Foreman MGG, Overland JE, Lorenzo ED, Bond NA, Aydin KY. 2011. Climate forcing and the California Current ecosystem. ICES J. Mar. Sci. 68(6):1199-1216.

King NJ, Priede IG. 2008. Coryphaenoides armatus, the Abyssal Grenadier: global distribution, abundance, and ecology as determined by baited landers. Am. Fish. Soc. Symp. 63:139-161.

Körtzinger A, Koeve W, Kähler P, Mintrop L. 2001. C:N ratios in the mixed layer during the productive season in the northeast Atlantic Ocean. Deep Sea Res. Part I Oceanogr. Res. Pap. 48(3):661-688.

Koslow JA. 1996. Energetic and life-history patterns of deep-sea benthic, benthopelagic and seamount-associated fish. J. Fish Biol. 49:54-74.

Kuhnz LA, Ruhl HA, Huffard CL, Smith KL Jr. 2014. Rapid changes and long-term cycles in the benthic megafaunal community observed over 24 years in the abyssal northeast Pacific. Prog. Oceanogr. 124:1-11.

Laptikhovsky V, Gaither MR, Black A. 2013. A pacific grenadier coryphaenoides acrolepis in the south-west atlantic and environmental changes in the falkland deep seas. Mar. Biodivers. Rec. 6:5.

Large PA, Hammer C, Bergstad OA, Gordon JDM, Lorance P. 2003. Deep-water fisheries of the northeast Atlantic: II. Assessment and management approaches. J. Northwest Atl. Fish. Sci. 31:151.

Leaman BM, Beamish RJ. 1984. Ecological and management implications of longevity in some northeast Pacific groundfishes. Bull. Int. North Pac. Fish. Comm. 42:85-97.

Leduc D, Rowden AA, Pilditch CA, Maas EW, Probert PK. 2013. Is there a link between deep-sea biodiversity and ecosystem function? Mar. Ecol. 34(3):334–344.

Lemon L. 2018. Population density and spatial distribution of abyssal epibenthic holothurians using fine scale time series imagery (2007-2017). SNS Master's Theses. 38. [Internet]. [cited 2019 Dec 8]. Available from: https://digitalcommons.csumb.edu/sns_theses/38

Lindegren M, Checkley DM, Ohman MD, Koslow JA, Goericke R. 2016. Resilience and stability of a pelagic marine ecosystem. Proc. R. Soc. B. 283(1822):20151931.

Mangubhai S, Saleh M, Muljadi A, Rhodes KL, Tjandra K. 2011. Do not stop: the importance of seamless monitoring and enforcement in an Indonesian marine protected area. J. Mar. Biol.

Matsui T, Kato S, Smith SE. 1991. Biology and potential use of Pacific grenadier, Coryphaenoides acrolepis, off California. Mar. Fish. Rev. 52(3):1-17.

Mazzini PLF, Barth JA, Shearman RK, Erofeev A. 2014. Buoyancy-driven coastal currents off Oregon during fall and winter. J. Phys. Oceanogr. 44(11):2854-2876.

McLeod E, Salm R, Green A, Almany J. 2009. Designing marine protected area networks to address the impacts of climate change. Front. Ecol. Environ. 7(7):362-370.

Meehl GA, Zwiers F, Evans J, Knutson T, Mearns L, Whetton P. 2000. Trends in extreme weather and climate events: Issues related to modeling extremes in projections of future climate change. Bull. Am. Meteorol. Soc. 81(3):427-436.

Melnychuk MC, Banobi JA, Hilborn R. 2013. Effects of management tactics on meeting conservation objectives for western North American groundfish fisheries. PloS One. 8(2):e56684.

Mendelssohn R, Schwing FB, Bograd SJ. 2003. Spatial structure of subsurface temperature variability in the California Current, 1950-1993. J. Geophys. Res. 108:3093.

Merrett NR, Haedrich RL. 1997. Deep-sea demersal fish and fisheries. Springer Sci. & Bus. Media. 23.

Mitchell JF. 1989. The "greenhouse" effect and climate change. Rev. Geophys. 27(1):115-139.

Mooney H, Zavaleta E. 2016. Ecosystems of California. Oakland (CA): University of California Press.

Moore JK, Lindsay K, Doney SC, Long MC, Misumi K. 2013. Marine ecosystem dynamics and biogeochemical cycling in the Community Earth System Model [CESM1 (BGC)]: Comparison of the 1990s with the 2090s under the RCP4. 5 and RCP8. 5 scenarios. J. Clim. 26(23):9291-9312.

Moore JE, Wallace BP, Lewison RL, Žydelis R, Cox TM, Crowder LB. 2009. A review of marine mammal, sea turtle and seabird bycatch in USA fisheries and the role of policy in shaping management. Mar. Policy. 33(3):435-451.

Morgan SG, Fisher JL. 2010. Larval behavior regulates nearshore retention and offshore migration in an upwelling shadow and along the open coast. Mar. Ecol. Prog. Ser. 404:109-126.

Morley S, Belchier M. 2002. Otolith and body size relationships in bigeye grenadier (Macrourus holotrachys) in CCAMLR subarea 48.3. CCAMLR Sci. 9:133-143.

Mueter FJ, Norcross BL. 2002. Spatial and temporal patterns in the demersal fish communities of the shelf and upper slope regions of the Gulf of Alaska. Fish. Bull. 10:559-581.

Mullin MM, Goetze E, Beaulieu SE, Lasker JM. 2000. Comparisons within and between years resulting in contrasting recruitment of Pacific hake (Merluccius productus) in the California Current System. Can. J. Fish. Aquat. Sci. 57:1434-1447.

National Oceanic and Atmospheric Administration. [date unknown]. Pacific whiting (hake). [Internet]. [cited 2019 October 25]. Available from: http://salishseanews.blogspot.com/2018/03/38-whiting-oil-safety-no-to-coast.html

Neat F, Burns F. 2010. Stable abundance, but changing size structure in grenadier fishes (Macrouridae) over a decade (1998–2008) in which deepwater fisheries became regulated. Deep Sea Res. Part I Oceanogr. Res. Pap. 57(3):434-440.

Nordquist M, Nandan SN, editors. 2011. United Nations Convention on the law of the sea 1982, Volume VII: a commentary. Martinus Nijhoff Publishers 7:222.

Norse EA, Brooke S, Cheung WW, Clark MR, Ekeland I, Froese R, Gjerde KM, Haedrich RL, Heppell SS, Morato T, Morgan LE. 2012. Sustainability of deep-sea fisheries. Mar. Policy. 36(2):307-320.

Nye JA, Link JS, Hare JA, Overholtz WJ. 2009. Changing spatial distribution of fish stocks in relation to climate and population size on the Northeast United States continental shelf. Mar. Ecol. Prog. Ser. 393:111-129.

Osmundson TW, Robert VA, Schoch CL, Baker LJ, Smith A, Robich G, Mizzan L, Garbelotto MM. 2013. Filling gaps in biodiversity knowledge for macrofungi: Contributions and assessment of an herbarium collection DNA barcode sequencing project. PLoS One. 8(4).

Overland JE, Alheit J, Bakun A, Hurrell JW, Mackas DL, Miller AJ. 2010. Climate controls on marine ecosystems and fish populations. J. Mar. Syst. 79(3-4):305-315.

Pankhurst NW, Munday PL. 2011. Effects of climate change on fish reproduction and early life history stages. Mar. Freshw. Res. 62(9):1015-1026.

Pansch C, Nasrolahi A, Appelhans YS, Wahl M. 2012. Impacts of ocean warming and acidification on the larval development of the barnacle Amphibalanus improvisus. J. Exp. Mar. Biol. Ecol. 420(421):48–55.

Patz JA, Grabow ML, Limaye VS. 2014. When it rains, it pours: future climate extremes and health. Ann. Glob. Health. 80(4):332-344.

Pauly D. 2009. Beyond duplicity and ignorance in global fisheries. Sci. Mar. 73(2):215–224.

Pecl GT, Araújo MB, Bell JD, Blanchard J, Bonebrake TC, Chen IC, Clark TD, Colwell RK, Danielsen F, Evengård B, Falconi L. 2017. Biodiversity redistribution under climate change: Impacts on ecosystems and human well-being. Science. 355(6332): p.eaai9214.

Pershing AJ, Mills KE, Record NR, Stamieszkin K, Wurtzell KV, Byron CJ, Fitzpatrick D, Golet WJ, Koob E. 2015. Evaluating trophic cascades as drivers of regime shifts in different ocean ecosystems. Philos. Trans. R. Soc. Lond. B Biol. Sci.

Peters RL. 1985. The greenhouse effect and nature reserves. Bioscience. 35(11):707-717.

Peterson WT, Emmett R, Goericke R, Venrick E, Mantyla A, Bograd SJ, Schwing FB et al. 2006. The state of the California Current, 2005-2006: warm in the north, cool in the south. Calif. Coop. Ocean. Fish. Invest. Rep. 47:30.

Petrik CM, Duffy-Anderson JT, Mueter F, Hedstrom K, Curchitser EN. 2015. Biophysical transport model suggests climate variability determines distribution of Walleye Pollock early life stages in the eastern Bering Sea through effects on spawning. Prog. Oceanogr. 138:459–474.

Pickett MH, Paduan JD. 2003. Ekman transport and pumping in the California Current based on the U.S. Navy's high- resolution atmospheric model (COAMPS). J. Geophys. Res. 108(C10): 3327.

Pikitch E, Santora C, Babcock EA, Bakun R, Bonfil DO, Conover P, Dayton PA, Doukakis P, Fluharty D, Heneman B, Houde ED. 2004. Ecosystem-based fishery management. Science. 305(5682):346-347.

Pinhassi J, Gómez-Consarnau L, Alonso-Sáez L, Sala MM, Vidal M, Pedrós-Alió C, Gasol JM. 2006. Seasonal changes in bacterioplankton nutrient limitation and their effects on bacterial community composition in the NW Mediterranean Sea. Aquat. Microb. Ecol. 44(3):241-252.

Powell TM, Lewis CVW, Curchitser EN, Haidvogel DB, Hermann AJ, Dobbins EL. 2006. Results from a three-dimensional, nested biological-physical model of the California Current System and comparisons with statistics from satellite imagery. J. Geophys. Res. Oceans. 111(7).

Priede IG, Deary AR, Bailey DM, Smith KL Jr. 2003. Low activity and seasonal change in population size structure of grenadiers in the oligotrophic abyssal Central North Pacific Ocean. J. Fish. Biol. 63:187-196.

Ragueneau O, Tréguer P, Leynaert A, Anderson RF, Brzezinski MA, DeMaster DJ, Dugdale RC, Dymond J, Fischer G, Francois R, Heinze C, Maier-Reimer E, Martin-Jézéquel V, Nelson DM, Quéguiner B. 2000. A review of the Si cycle in the modern ocean: Recent progress and missing gaps in the application of biogenic opal as a paleoproductivity proxy. Glob. Planet. Change. 26(4):317–365.

Ramirez-Llodra E, Brandt A, Danovaro R, De Mol B, Escobar E, German C, Levin L, Arbizu P, Menot L, Buhl-Mortensen P, Narayanaswamy BE, Smith CR, Tittensor DP, Tyler PA, A. Vanreusel A, Vecchione M. 2010. Deep, diverse and definitely different: Unique attributes of the world's largest ecosystem. Biogeosciences. *7*(9):2851-2899.

Recent Global Warming. 2002. [Internet]. [cited 2018 July 5]. Available from: http://earthguide.ucsd.edu/virtualmuseum/climatechange2/08 1.shtml

Reuter RF, Conners ME, DiCosimo J, Gaichas S, Ormseth O, TenBrink TT. 2010. Managing non-target, data-poor species using catch limits: lessons from the Alaskan groundfish fishery. Fish. Manag. Ecol. 17(4):323-335.

Rijnsdorp AD, Peck MA, Engelhard GH, Mollmann C, Pinnegar JK. 2009. Resolving the effect of climate change on fish populations. ICES J. Mar. Sci. 66(7):1570-1583.

Ripple WJ, Estes JA, Beschta RL, Wilmers CC, Ritchie EG, Hebblewhite M, Berger J, Elmhagen B, Letnic M, Nelson MP, Schmitz OJ. 2014. Status and ecological effects of the world's largest carnivores. Science. 343(6167):1241484.

Roberts CM. 2002. Deep impact: the rising toll of fishing in the deep sea. Trends Ecol. Evol. 17(5):242-245.

Rosenzweig C, Karoly D, Vicarelli M, Neofotis P, Wu Q, Casassa G, Menzel A, Root TL, Estrella N, Seguin B, Tryjanowski P, Chunzhen L, Rawlins S, Imeson A. 2008. Attributing physical and biological impacts to anthropogenic climate change. Nature. 453(7193):353-357.

Ruhl HA. 2008. Community change in the variable resource habitat of the abyssal Northeast Pacific. Ecology. 89(4):991-1000.

Schlining, BM, Stout NJ. 2006. MBARI's video annotation and reference system. Oceans. (IEEE, 2006):1-5.

Schneider SH. 1989. The greenhouse effect: science and policy. Science. 243(4892):771-781.b.

Schneider B, Bopp L, Gehlen M, Segschneider J, Frölicher TL, Cadule P, Friedlingstein P, Doney SC, Behrenfeld MJ, Joos F. 2008. Climate-induced interannual variability of

marine primary and export production in three global coupled climate carbon cycle models. Biogeosciences. 5(2):597–614.

Sherman, AD, Smith KL Jr. 2009. Deep-sea benthic boundary layer communities and food supply: A long-term monitoring strategy. Deep Sea Res. Part II Top. Stud. Oceanogr. 56(19-20): 1754-1762.

Shotton R. 2005. C3. Review of the state of world marine fishery resources. Deepwater fisheries. FAO Fish. Aquat. Tech. Pap. 188.

Smith KL Jr., Sherman AD, McGill PR, Henthorn RG, Ferreira J, Huffard CL. 2017. Evolution of monitoring an abyssal time-series station in the northeast Pacific over 28 years. Oceanogr. 30(4):72-81.

Smith KL, Ruhl HA, Kahru M, Huffard CL, Sherman AD. 2013. Deep ocean communities impacted by changing climate over 24 y in the abyssal northeast Pacific Ocean. Proc. Natl. Acad. Sci. U.S.A. 110(49):19838-19841.

Smith KL Jr, Baldwin RJ, Ruhl HA, Kahru M, Mitchell BG, Kaufmann RS. 2006. Climate effect on food supply to depths greater than 4,000 meters in the northeast Pacific. Limnol. Oceanogr. 51(1): 166-176.

Smith KL Jr, Huffard CL, Ruhl HA. 2020. Thirty-year time series study at a station in the abyssal NE Pacific: An introduction. Deep Sea Res. Part II Top. Stud. Oceanogr. 104764. Smith KL Jr, Kaufmann RS, Baldwin RJ. 1994. Coupling of near-bottom pelagic and benthic processes at abyssal depths in the eastern North Pacific Ocean. Limnol. Oceanogr. 39(5):1101-1118.

Smith KL Jr, Kaufmann RS, Wakefield WW. 1993. Mobile megafaunal activity monitored with a timelapse camera in the abyssal North Pacific. Deep Sea Res. Part I Oceanogr. Res. Pap. 40: 2307-2324.

Smith KL Jr, Whit GA, Laver MB, McConnaughey RR, Meador JP. 1979. Free vehicle capture of abyssopelagic animals. Deep-Sea Res. Oceanogr., A. 26(1):57-64.

Smith PE. 1985. A case history of an anti-El Niño to El Niño transition on plankton and nekton distribution and abundances. In: Wooster WS, Fluharty DL, editors. El Niño north: Niño effects in the eastern subarctic Pacific Ocean. Seattle (WA): Washington Sea Grant Program, University of Washington. 121-142.

Smith PE. 1985. Year-class strength and survival of O-group clupeoids. Can. J. Fish. Aquat. Sci. 42(S1):s69-s82.

Stauffer GD. 1985. Biology and life history of the coastal stock of Pacific whiting, Merluccius productus. Mar. Fish. Rev. 47(2): 2-7.

Stein DL, Pearcy WG. 1982. Aspects of reproduction, early life history, and biology of macrourid fishes off Oregon, USA. Deep-Sea Res. Oceanogr., A. 29:1313-1329.

Steinberg DK, Lomas MW, Cope JS. 2012. Long-term increase in mesozooplankton biomass in the Sargasso Sea: Linkage to climate and implications for food web dynamics and biogeochemical cycling. Glob. Biogeochem. Cycles. 26(1).

Street-Perrott FA, Barker PA, Leng MJ, Sloane HJ, Wooller MJ, Ficken KJ, Swain DL. 2008. Towards an understanding of late quaternary variations in the continental biogeochemical cycle of silicon: multi-isotope and sediment-flux data for Lake Rutundu, Mt Kenya, East Africa, since 38 ka BP. J. Quat. Sci. 23(4):375-387.

Strub PT, James C, Thomas AC, Abbott MR. 1990. Seasonal and nonseasonal variability of satellite-derived surface pigment concentration in the California Current. J. Geophys. Res. 92:1501-1530.

Suikkanen S, Pulina S, Engström-Öst J, Lehtiniemi M, Lehtinen S, Brutemark A. 2013. Climate change and eutrophication induced shifts in northern summer plankton communities. PLoS One. 8(6).

Sukresno B. 2010. Empirical orthogonal functions (EOF) analysis of SST variability in Indonesian water concerning with ENSO and IOD. ISPRS Archives. 38(8):116-121.

Sumida BY, Moser HG. 1980. Food and feeding of Pacific hake larvae, Merluccius productus, off southern California and northern Baja California. Work. 161.

Therriault T, Park C, Rice J. 2016. C. North Pacific Ocean.

Tréguer PJ, De La Rocha CL. 2013. The world ocean silica cycle. Annu. Rev. Mar. Sci. 5:477-501.

UNEP. 2006. Ecosystems and biodiversity in deep waters and high seas. UNEP. Reg. Seas Rep. Stud. 178:1-58. [Internet]. [cited 2019 May 11]. Available from: http://www.unep.org/pdf/EcosystemBiodiversity_DeepWaters_20060616.pdf

Vanni MJ. 1987. Effects of food availability and fish predation on a zooplankton community. Ecol. Monogr. 57(1):61-88.

Velaquéz AV. 2003. Reproductive strategies of the spiny lobster Panulirus interruptus related to the marine environmental variability off central Baja California, Mexico: Management implications. Fish. Res. 65:123–135.

Walker HJ Jr, Watson W, Barnett AM. 1987. Seasonal occurrence of larval fishes in the nearshore Southern California Bight off San Onofre, California. Estuar. Coast. Shelf Sci. 25(1):91-109.

Wallace JR. Aquatic vertebrate carrion decomposition. In: Benbow ME, Tomberlin JK, Tarone AM, editors. 2015. Carrion ecology, evolution, and their applications. Boca Raton (FL): CRC Press. P. 247-272.

Wegner G, Damm U, Purps M. 2003. Physical influences on the stock dynamics of plaice and sole in the North Sea. Sci. Mar. 67:219–234.

Wilson RR Jr, Attia P. 2003. Interrelationships of the subgenera of Coryphaenoides (Teleostei: Gadiformes: Macrouridae): synthesis of allozyme, peptide mapping, and DNA sequence data. Mol. Phylogenetics Evol. 27(2):343-347.

Wilson RR, Waples RS. 1983. Distribution, morphology, and biochemical genetics of Coryphaenoides armatus and C. yaquinae (Pisces: Macrouridae) in the central and eastern North Pacific. Deep-Sea Res. Oceanogr., A. 30(11):1127-1145.

Wolter K, Timlin MS. 2011. El Niño/Southern Oscillation behaviour since 1871 as diagnosed in an extended multivariate ENSO index (MEI. ext). Int. J. Climatol. 31(7):1074-1087.

Yoklavich MM, Greene HG, Cailliet GM, Sullivan DE, Lea RN, Love MS. 2000. Habitat associations of deep-water rockfishes in a submarine canyon: an example of a natural refuge. Fish. Bull. (Wash. D.C.). 98:625-641.

Young JW, Hunt BPV, Cook TR, Llopiz JK, Hazen EL, Pethybridge HR, Ceccarelli D, Lorrain A, Olson RJ, Allain V, Menkes C, Patterson T, Nicol S, Lehodey P, Kloser RJ, Arrizabalaga H, Choy CA. 2015. The trophodynamics of marine top predators: Current knowledge, recent advances and challenges. Deep Sea Res. Part II Top. Stud. Oceanogr. 113:170–187.

Zeidberg LD, Robison BH. 2007. Invasive range expansion by the Humboldt squid, Dosidicus gigas, in the eastern North Pacific. Proc. Natl. Acad. Sci. U.S.A. 104:12948-12950.

Zeppilli D, Pusceddu A, Trincardi F, Danovaro R. 2016. Seafloor heterogeneity influences the biodiversity-ecosystem functioning relationships in the deep sea. Sci. Rep. 6.

APPENDIX A

LIST OF HYPOTHESIS

Question #	Question Section			Question &	Hypotheses	Statistical Tests	
		Has Me	acrourid species complex' populati	on structure change condi	Mann-Kendall		
1	A	<i>Ho:</i> Wi	th increasing changes in climate a	reasing changes in climate and surface ocean conditions over time, there is no significant change in Macrourid species complex' abundance.			
		<i>Ha:</i> With i	ncreasing changes in climate and	surface ocean condit complex' a	ions over time, there is a significant change in Macrourid species bundance.	Spearman Rank	
	Fich Variable		Cracios Included	Data Source	Climate /Surface Condition Variables	Time Carios	

Fish Variable	Species Included	Data Source		Climate/Surface Condition Variables					
Macrourid species complex' abundance (# ind. /hr/ week)	Coryphaenoides armatus Corphaenoides leptolepis Coryphaenoides yaquinae	Sta. M	POC	C_N	MEI	SST	NPP	2011-2017	

Question #	Question Section				Statistical Tests					
		Has M	acrourid sp.' population structure	an conditions?	Man	n-Kendall				
1	В	Ho: W	th increasing changes in climate a	Cross Corre	Cross Correlation Analysis					
	Ha: With increasing changes in climate and surface ocean conditions over time, there is a significant change in Macrourie species' biomass.						e in Macrourid	Spear	Spearman Rank	
Fish Variables			Species Included	Data Source	Climate/Surface Condition Variables					Time-Series
Macrourid biomass (kg per ha/der) (Regional)		'S	Nezumia stelgidolepis Albatrossia pectoralis Coryphaenoides acrolepis Nezumia liolepis Coryphaenoides cinereus Coelorinchus scaphopsis Malacocephalus laevis Grenadier unident.	Mined	POC	C_N	MEI	SST	NPP	2011-2017

1 C Has Macrourid sp.' population structure changed over time in relation to changing climate and surface-ocean conditions? Ho: With increasing changes in climate and surface ocean conditions over time, there is no significant change in Macrourid species CPUE. Crophysical CPUE Ha: With increasing changes in climate and surface ocean conditions over time, there is a significant change in Macrourid species CPUE. Crophysical CPUE	Question # Question	n Section	ction Question & Hypotheses								
1 C Ho: With increasing changes in climate and surface ocean conditions over time, there is no significant change in Macrourid species' CPUE. Crossing changes in climate and surface ocean conditions over time, there is a significant change in Macrourid reserves CPUE.			Has Macrourid sp.' population structure chang	s Macrourid sp.' population structure changed over time in relation to changing climate and surface-ocean conditions?							
Ha: With increasing changes in climate and surface ocean conditions over time, there is a significant change in Macrourid	1 0	<i>Но</i> : V С	Ho: With increasing changes in climate and su	Cross Correlation Analysis							
species CPOE.			Ha: With increasing changes in climate and su	surface ocean condi species' Cf	itions over time, there is a significant change in Macrourid PUE.	Spearman Rank					

Fish Variable	Species Included	Data Source		Climate/Surface Condition Variables				
Macrourid CPUE (num per ha/der) (Regional)	Nezumia stelgidolepis Albatrossia pectoralis Coryphaenoides acrolepis Nezumia liolepis Coryphaenoides cinereus Coelorinchus scaphopsis Malacocephalus laevis Grenadier unident.	Mined	POC	C_N	MEI	SST	NPP	2011-2017

Question #	Question Section			Question & H	lypotheses				Stati	stical Tests
		Has M	acrourid sp.' population structure	changed over time in	relation to cha	nging climate o	and surface-oce	an conditions?	Mai	nn-Kendall
1	D	<i>Ho:</i> Wi	th increasing changes in climate and surface ocean conditions over time, there is no significant change in Macrourid species' body lengths.						Cross Corr	elation Analysis
		Ha: W	ith increasing changes in climate	and surface ocean cor species' boo	nditions over tii dy lengths.	me, there is a s	ignificant chan	ge in Macrourid	Spea	rman Rank
	Fish Variable		Species Included	Data Source Climate/Surface Condition Variables					Time-Series	

Macrourid length (cm) (Regional)	Nezumia stelgidolepis Albatrossia pectoralis Coryphaenoides acrolepis Nezumia liolepis	Mined	POC	C_N	MEI	SST	NPP	2011-2017
-------------------------------------	---	-------	-----	-----	-----	-----	-----	-----------

Question #	Question Section			Question &	Hypotheses				Statis	tical Tests
		Has Ma	acrourid sp.' population structure	changed over time in	n relation to cho	nging climate c	nd surface-oce	an conditions?	Man	n-Kendall
1	E	Ho: With	increasing changes in climate and	in Macrourid sp	.' Cross Corr	elation Analysis				
		Ha: With	Ha: With increasing changes in climate and surface ocean conditions over time, there is a significant change in Macrourid sp larval abundances.							
	Fish Variables		Species Included	Data Source		Climate/Su	face Conditio	on Variables		Time-Series
Macro	Macrourid larval abundance (10 m ²) (Regional)		Nezumia sp. Albatrossia pectoralis Coryphaenoides acrolepis Macrouridae sp.	Mined	POC	C_N	MEI	SST	NPP	2011-2017

Question #	Question Section		Question & Hypo	theses	Statistical Tests				
	A	Has Merluccius productus' population structure	lerluccius productus' population structure changed over time in relation to changing climate and surface-ocean conditions?						
2		Ho: With increasing changes in climate and so	surface ocean condition productus' bior	ns over time, there is no significant change in <i>Merluccius</i> nass.	Cross Correlation Analysis				
		Ha: With increasing changes in climate and s	surface ocean conditic productus' bior	ins over time, there is a significant change in <i>Merluccius</i> nass.	Spearman Rank				

Fish Variables	Species Included	Data Source		Climate/Surface Condition Variables					
<i>M. productus</i> biomass (kg per ha/der) (Regional & Local)	Merluccius productus	Mined	POC	C_N	MEI	SST	NPP	2011-2017	

Question #	Question Section			Question &	Hypotheses				Statis	tical Tests
		Has Merl	uccius productus' population struct	ture changed over tin	ne in relation to	changing clima	te and surface of	ocean condition	s? Man	n-Kendall
2	В	Ho: Wi	th increasing changes in climate a	e in Merluccius	Cross Corre	Cross Correlation Analysis				
		Ha: W	Ha: With increasing changes in climate and surface ocean conditions over time, there is a significant change in Merluccius productus' CPUE.							
							<u> </u>		-	
	Fish Variables		Species Included	Data Source		Climate/Su	rface Conditio	n Variables		Time-Series
M. (1 (F	<i>M. productus</i> CPUE (num per ha/der) (Regional & Local)		Merluccius productus	Mined	POC	C_N	MEI	SST	NPP	2011-2017

Question #	Question Section			Question &	Hypotheses				Statis	tical Tests
		Has Merlu	? Man	n-Kendall						
2 C Ho: With increasing changes in climate and surface ocean conditions over time, there is no significant change in Merluccius productus' body lengths.								Cross Corr	Cross Correlation Analysis	
	Ha: With increasing changes in climate and surface ocean conditions over time, there is a significant change in Merluccius productus' body lengths.							Spear	Spearman Rank	
	Fish Variable		Species Included	Data Source		Climate/Su	rface Condition	on Variables		Time-Series
M. pr	roductus length (Regional)	(cm)	Merluccius productus	Mined	POC	C_N	MEI	SST	NPP	2011-2017

Question #	Question Section	Question & Hypotheses	Statistical Tests
2	Has Merluccius productus' population structure changed over	Has Merluccius productus' population structure changed over time in relation to changing climate and surface ocean conditions?	Mann-Kendall
	D	Ho: With increasing changes in climate and surface ocean conditions over time, there is no significant change in Merluccius productus' larval abundance.	Cross Correlation Analysis
		Ha: With increasing changes in climate and surface ocean conditions over time, there is a significant change in Merluccius productus' larval abundance.	Spearman Rank

Fish Variables	Species Included	Species Included Data Source		Climate/Surface Condition Variables						
<i>M. productus</i> larval abundance (10 m ²) (Regional)	Merluccius productus	Mined	POC	C_N	MEI	SST	NPP	2011-2017		

Question #	Question Section	Question & Hypotheses	Statistical Tests
3		Has the depth range of Macrourid species changed over time with changing climate and surface ocean conditions?	Mann-Kendall
	A	Ho: With increasing changes in climate and surface ocean conditions over time, there is no significant change in Macron A species' depth distributions.	
		Ha: With increasing changes in climate and surface ocean conditions over time, there is a significant change in Macrourid species' depth distributions.	Spearman Rank

Fish Variable	Species Included	Data Source		Climate/Su	rface Condition	on Variables		Time-Series
Macrourid depth (m) (Regional)	Nezumia stelgidolepis Albatrossia pectoralis Coryphaenoides acrolepis Nezumia liolepis Coryphaenoides cinereus Coelorinchus scaphopsis Malacocephalus laevis Grenadier unident.	Mined	POC	C_N	MEI	SST	NPP	2011-2017

Question #	Question Section	Question & Hypotheses	Statistical Tests
4		Has the depth range of Merluccius productus changed over time with changing climate and surface ocean conditions?	Mann-Kendall
	A	Ho: With increasing changes in climate and surface ocean conditions over time, there is no significant change in Merluccius productus' depth distribution.	Cross Correlation Analysis
		Ha: With increasing changes in climate and surface ocean conditions over time, there is a significant change in Merluccius productus' depth distribution.	Spearman Rank

Fish Variables	Species Included Data Source			Climate/Surface Condition Variables						
<i>M. productus</i> depth (m) (Regional & Local)	Merluccius productus	Mined	POC	C_N	MEI	SST	NPP	2011-2017		

APPENDIX B

RESULTS TABLES

MACROURIDS (GROUPED)

	Variables		Cross Correlation			Spearman Rank		Mann-Kendall Trend Tests (Time Series Trend)				
Fish Variable	Scale	Environmental Variable	Correlation	Lag (# weeks)	Direction (+/-)	rho	P-Value	Hypothesis	Trend	S (Test Statistic)	Trend Direction	P-Value
Manalia		POC	-0.23	6	(-)	-0.22	0.001					
	T 1	C_N	-0.31	67	(-)	-0.32	0.000					
abundance	(Sta, M)	SST	0.21	22	(+)	0.20	0.001	Ha	Yes	8590	Increasing	< 0.0001
	()	NPP	-0.36	50	(-)	-0.35	0.000					
		MEI	0.39	56	(+)	0.40	0.000					
		POC	0.33	0	(+)	0.27	0.042					
		C_N	0.24	1	(+)	0.29	0.043				NA	NA
Macrourid biomass	Regional	SST	-0.15	2	(-)	-0.15	0.181	H0 No	No	NA		
		NPP	-0.23	3	(-)	-0.28	0.010					
		MEI	0.24	102	(+)	0.21	0.073					
	Regional	POC	-0.29	14	(-)	-0.24	0.041	H0 No		NA	NA	NA
		C_N	0.26	3	(+)	0.25	0.084		No			
Macrourid CPUE		SST	-0.13	3	(-)	-0.13	0.244					
		NPP	-0.16	3	(-)	-0.20	0.072					
		MEI	0.14	50	(+)	0.15	0.181					
		POC	0.27	3	(+)	0.23	0.099					
		C_N	-0.33	2	(-)	-0.35	0.020					
Macrourid depth	Regional	SST	0.19	36	(+)	0.14	0.254	H0	No	NA	NA	NA
		NPP	-0.20	25	(-)	-0.23	0.048					
		MEI	0.15	8	(+)	0.15	0.211					
		POC	0.27	0	(+)	0.04	0.801					
		C_N	-0.48	39	(-)	-0.53	0.000					0.0023
Macrourid length	Regional	SST	0.33	30	(+)	0.40	0.001	На	Yes	640	Increasing	
		NPP	-0.52	56	(-)	-0.58	0.000					
		MEI	0.34	23	(+)	0.35	0.002					

Table #1. Table of statistical analyses results for relationships between Macrourid spp.' variables and environmental variables. Gray and bold cells represent most significant findings.

MERLUCCIUS PRODUCTUS

	Variables		Cross Correlation			Spearman Rank		Mann-Kendall Trend Tests (Time Series Trend)					
Fish Variable	Scale	Environmental Variable	Correlation	Lag (# weeks)	Direction (+/-)	rho	P-Value	Hypothesis	Trend	S (Test Statistic)	Trend Direction	P-Value	
		POC	-0.25	4	(-)	-0.21	0.101						
M. productus biomass		C_N	0.34	2	(+)	0.32	0.024					1	
	Regional	SST	-0.19	38	(-)	-0.17	0.114	H0	No	NA	NA	NA	
		NPP	0.34	12	(+)	0.35	0.001						
		MEI	-0.20	4	(-)	-0.23	0.033						
		POC	0.31	16	(+)	0.30	0.013						
		C_N	0.35	18	(+)	0.33	0.007						
M. productus CPUE	Regional	SST	-0.27	39	(-)	-0.27	0.012	H0	No	NA	NA	NA	
		NPP	0.30	7	(+)	0.34	0.002						
		MEI	-0.26	1	(-)	-0.29	0.008						
		POC	0.25	0	(+)	0.07	0.839						
		C_N	-0.53	3	(-)	-0.55	0.171	-					
M. productus biomass	Local	SST	0.73	6	(+)	0.75	0.000	На	Yes	63	Increasing	0.017	
*		NPP	-0.53	2	(-)	-0.68	0.003						
		MEI	0.68	15	(+)	0.76	0.000						
	Local	POC	0.53	7	(+)	0.47	0.070	H0	No	o NA	NA	NA	
		C_N	0.45	0	(+)	0.53	0.148						
M. productus CPUE		SST	0.83	6	(+)	0.86	< 2.2e-16						
		NPP	-0.67	2	(-)	-0.80	0.000						
		MEI	0.48	15	(+)	0.55	0.018						
		POC	0.35	14	(+)	0.30	0.010				NA	NA	
		C_N	0.34	39	(+)	0.31	0.027			NA			
M. productus depth	Regional	SST	0.22	1	(+)	0.30	0.006	H0	No				
		NPP	0.24	16	(+)	0.26	0.017						
M. productus depth		MEI	-0.13	0	(-)	-0.15	0.180					i i	
		POC	-0.58	8	(-)	-0.58	0.016						
		C_N	0.57	3	(+)	0.90	0.005						
M. productus depth	Local	SST	-0.32	6	(-)	-0.30	0.232	H0	No	NA	NA	NA	
		NPP	0.20	12	(+)	0.26	0.294						
		MEI	0.10	6	(+)	0.07	0.794	-					
		POC	-0.39	4	(-)	-0.45	0.000						
		C_N	0.24	2	(+)	0.22	0.125						
M. productus length	Regional	SST	-0.25	4	(-)	-0.28	0.010	H0	No	NA	NA	NA	
		NPP	-0.22	15	(-)	-0.24	0.028						
		MEI	0.23	52	(+)	0.25	0.024	1					
		POC	-0.32	5	(-)	-0.20	0.168						
		C_N	0.64	4	(+)	0.58	0.000						
M. productus larval	Regional	SST	0.58	14	(+)	0.55	0.000	H0	No	NA	NA	NA	
abundance		NPP	-0.39	2	(-)	-0.45	0.000						
		MEI	-0.18	0	(-)	-0.17	0.198						

Table #2. Table of statistical analyses results for relationships between *M. productus* 'variables and environmental variables. Gray and bold cells represent most significant findings.

APPENDIX C

RESULTS PLOTS

VARIABLES OVER TIME

Environmental Variables



Plot 1. Weekly averaged POC flux (mg C/m²/d), at Station M, during the 2011-2017 study period.



Plot 2. Weekly averaged carbon to nitrogen ratios, at Station M, during the 2011-2017 study period.


Plot 3. Weekly averaged SST (°C), for the California coast, during the 2011-2017 study period.



Plot 4. Weekly averaged NPP (mg C/m2/d), at Station M, during the 2011-2017 study period.



Plot 5. Weekly averaged MEI (σ), for the California coast, during the 2011-2017 study period.

Macrourid Variables



Plot 6. Ranked weekly averaged Macrourid species complex' abundance (# indiv./hr), at Station M, during the 2011-2017 study period.



Plot 7. Ranked weekly averaged regional Macrourid spp.' biomass (CPUE kg/ha), for the California coast, during the 2011-2017 study period.



Plot 8. Ranked weekly averaged regional Macrourid spp.' CPUE (num/ha), for the California coast, during the 2011-2017 study period.



Plot 9. Ranked weekly averaged regional Macrourid spp.' length (cm), for the California coast, during the 2011-2017 study period.



Plot 11. Ranked weekly averaged regional Macrourid spp.' depth (m), for the California coast, during the 2011-2017 study period.



Merluccius Productus Variables

Plot 12. Ranked weekly averaged regional *M. productus* biomass (CPUE kg/ha), for the California coast, during the 2011-2017 study period.



Plot 13. Ranked weekly averaged *M. productus* biomass (CPUE kg/ha), localized within 100 nmi of Station M, during the 2011-2017 study period.



Plot 14. Ranked weekly averaged regional *M. productus* CPUE (num/ha), for the California coast, during the 2011-2017 study period.



Plot 15. Ranked weekly averaged *M. productus* CPUE (num/ha), localized within 100 nmi of Station M, during the 2011-2017 study period.



Plot 16. Ranked weekly averaged regional *M. productus* length (cm), for the California coast, during the 2011-2017 study period.



Plot 17. Ranked weekly averaged regional *M. productus* larval abundance $(10m^2)$, for the California coast, during the 2011-2017 study period.



Plot 18. Ranked weekly averaged regional *M. productus* depth (m), for the California coast, during the 2011-2017 study period.



Plot 19. Ranked weekly averaged *M. productus* depth (m), localized within 100 nmi of Station M, during the 2011-2017 study period.

CROSS CORRELATION

SCATTER PLOT





lag-number of weekly increments







Lag(Weeks)





Lag (Weeks)







lag-number of weekly increments





lag-number of weekly increments







